

## Very strong atmospheric methane growth in the four years 2014 - 2017: Implications for the Paris Agreement

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### Key Points

1. Atmospheric methane is rising. Its carbon isotopic ratio has become more depleted in C-13.
2. The possible causes of the change include increasing emissions, with changing relative proportions of source inputs, or a decline in methane destruction, or both.
3. If this rise continues, there are significant consequences for the UN Paris Agreement.

### Key Words

Atmospheric methane, Methane isotopes, Wetland methane emissions, Fossil fuel methane emissions, OH destruction of methane, Paris Agreement

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## Abstract

*Atmospheric methane grew very rapidly in 2014 ( $12.7 \pm 0.5$  ppb/yr), 2015 ( $10.1 \pm 0.7$  ppb/yr), 2016 ( $7.0 \pm 0.7$  ppb/yr) and 2017 ( $7.7 \pm 0.7$  ppb/yr), at rates not observed since the 1980s. The increase in the methane burden began in 2007, with the mean global mole fraction in remote surface background air rising from about 1775 ppb in 2006 to 1850 ppb in 2017. Simultaneously the  $^{13}\text{C}/^{12}\text{C}$  isotopic ratio (expressed as  $\delta^{13}\text{C}_{\text{CH}_4}$ ) has shifted, in a new trend to more negative values that have been observed worldwide for over a decade. The causes of methane's recent mole fraction increase are therefore either a change in the relative proportions (and totals) of emissions from biogenic and thermogenic and pyrogenic sources, especially in the tropics and sub-tropics, or a decline in the atmospheric sink of methane, or both. Unfortunately, with limited measurement data sets, it is not currently possible to be more definitive. The climate warming impact of the observed methane increase over the past decade, if continued at  $>5$  ppb/yr in the coming decades, is sufficient to challenge the Paris Agreement, which requires sharp cuts in the atmospheric methane burden. However, anthropogenic methane emissions are relatively very large and thus offer attractive targets for rapid reduction, which are essential if the Paris Agreement aims are to be attained.*

## Plain Language Summary

The rise in atmospheric methane ( $\text{CH}_4$ ), which began in 2007, accelerated in the past four years. The growth has been worldwide, especially in the tropics and northern mid-latitudes. With the rise has come a shift in the carbon isotope ratio of the methane. The causes of the rise are not fully understood, and may include increased emissions and perhaps a decline in the destruction of methane in the air. Methane's increase since 2007 was not expected in future greenhouse gas scenarios compliant with the targets of the Paris Agreement, and if the increase continues at the same rates it may become very difficult to meet the Paris goals. There is now urgent need to reduce methane emissions, especially from the fossil fuel industry.

## 1. Introduction

Methane is the second most important anthropogenic greenhouse gas (Myhre et al., 2013; Etminan et al., 2016; Allen et al., 2018). In the 1990s, the atmospheric methane burden trended towards equilibrium, which it reached by the end of the 20<sup>th</sup> century (Dlugokencky et al., 2011), with little or no growth in its atmospheric burden in the early years of this century. In 1984, the first year with detailed records, the global annual average atmospheric mole fraction of methane in the remote marine boundary layer was 1645 ppb. In 2006, just before the recent growth phase began, it was about 1775 ppb. This grew rapidly to an annual global mean of 1850 ppb in 2017, a total rise of about 75 ppb in the 2007-2017 period. The increase is continuing. The growth in the mole fraction of methane in Arctic air was strong in late 2007, but since then in many years the strongest growth has been in the tropics and subtropics of the Northern and Southern Hemispheres (Figures 1 and 2). Overall, since 2007, methane growth has been sustained globally. Pre-industrial methane was about 720 ppb around 1750 C.E. (Etheridge et al., 1998); thus the decade of recent growth is equivalent to more than 6% of total growth to date since industrialization began.

Since 2007, atmospheric methane's changing carbon isotope ratio (expressed as  $\delta^{13}\text{C}_{\text{CH}_4}$ ) implies there has also been a significant change in the balance of sources and sinks. In contrast to the 1980s, and indeed the past two centuries, when  $\delta^{13}\text{C}_{\text{CH}_4}$  showed a sustained shift to more positive values indicative of gas leaks and coal emissions (Lowe et al., 1994; Rice et al., 2016), the present rise has been accompanied by a sustained shift in  $\delta^{13}\text{C}_{\text{CH}_4}$  to more negative values (Figure 3) (Schaefer et al. 2016; Nisbet et al., 2016). Several hypotheses, not mutually exclusive, are possible to explain the changes. They are summarized here and then discussed in detail below (Section 6):

1. An increase has occurred in isotopically very negative biogenic emissions, whether from wetlands or ruminants or waste, or all of these. If so, an increase in the proportion of global emissions from microbial sources may have driven both the increase in the methane burden and the shift in  $\delta^{13}\text{C}_{\text{CH}_4}$  (Schaefer et al., 2016; Nisbet et al., 2016).
2. *Or* strong rise in methane emissions from the use of natural gas and oil has taken place (Hausmann et al., 2016). Fossil fuel methane emissions are mostly somewhat more positive in  $\delta^{13}\text{C}_{\text{CH}_4}$  than -47‰ to -53‰. Thus this hypothesis is only consistent with the observed isotopic shift if either: a) the new fossil fuel emissions have  $\delta^{13}\text{C}_{\text{CH}_4}$  markedly more negative than -47‰; or b) if there has been a concurrent decline in a source of much more  $^{13}\text{C}$  rich emissions, such as from biomass burning (Worden et al., 2017); or c) both changes have occurred. Note that it is possible that both hypotheses 1) and 2) may be valid: that both microbial and fossil fuel emissions have increased. This would explain the observations provided the increase in microbial emissions is sufficiently larger than that in fossil fuels, so that the bulk  $\delta^{13}\text{C}_{\text{CH}_4}$  value of the total source has become more negative.
3. *Or* the oxidative capacity – the cleansing power - of the atmosphere has declined, and hence the destruction of methane has slowed. A change in methane destruction has strong isotopic impact. If this hypothesis is correct, total emissions may have changed little or (less likely) even decreased, if the isotopic shift has been caused by a reduction in the OH sink (Turner et al., 2017; Rigby et al., 2017). However, Naus et al. (2019), using output from 3D model simulations, pointed out the limitations of two-box modelling. They found only small differences in OH anomalies (up to 1.3%, averaged over 1994-2015) relative to the full uncertainty envelope (5-8%).

Although methane's apparent equilibration in the early years of this century was perhaps only a temporary pause in the human-induced secular increase in atmospheric methane (Bousquet et al., 2006), the renewed strong methane growth that began in 2007 (Nisbet et al., 2014;

2016) was so unexpected that it was not considered in pathway models preparatory to the Paris Agreement (Moss et al., 2008; Meinshausen et al., 2011; Rogelj et al., 2012; Collins et al., 2013). The current growth has now lasted over a decade. If growth continues at similar rates through subsequent decades, evidence presented here demonstrates that the extra climate warming impact of the methane can significantly negate or even reverse progress in climate mitigation from reducing CO<sub>2</sub> emissions. This will challenge efforts to meet the target of the 2015 UN Paris Agreement on Climate Change, to limit climate warming to 2°C.

## 2. Methods

This paper focuses on direct *in situ* measurements of air. Air samples from remote sites reported here come from two groups of sources. Weekly air samples are collected in flask pairs at ~60 sampling sites in the Cooperative Global Air Sampling Network of the National Oceanic and Atmospheric Administration (NOAA), USA, a component of their Global Greenhouse Gas Reference Network. Air sampling, analysis, and data smoothing methods for methane are described in Dlugokencky et al. (1994). All methane measurements are reported on the WMO X2004A scale (Dlugokencky et al., 2005; [www.esrl.noaa.gov/gmd/cc1/ch4\\_scale.html](http://www.esrl.noaa.gov/gmd/cc1/ch4_scale.html)) in mole fraction units, nmol mol<sup>-1</sup> dry air, or parts per billion, abbreviated as “ppb”. Uncertainties range from ~±3 ppb in the early 1980s to ~±1 ppb in 2017 as measurement precisions have improved. A subset of NOAA samples is analyzed for δ<sup>13</sup>C<sub>CH4</sub>; analysis methods are described in Miller et al. (2002); see also White et al. (2017).

The second data set is from the flask sampling program of the Greenhouse Gas Laboratory at Royal Holloway, University of London (RHUL) and the Norwegian Institute for Air Research (NILU) as well as from RHUL’s cavity-ring-down absorption spectrometer operating on Ascension Island, UK and NILU’s measurement at Zeppelin, Svalbard (Spitsbergen). For methods, see Nisbet et al. (2016) and Fisher et al. (2006 and 2017). The RHUL time series were collected under a variety of projects supported by the European Union and the UK Natural Environment Research Council (NERC), and most recently the NERC “MOYA” Global Methane Budget project, with help from Environment Canada and the University of Heidelberg (UHEI; samples from Alert), NILU (Zeppelin), the UK Met Office (Ascension) and the South African Weather Service (Cape Point).

All air samples from remote sites are analyzed in the laboratories in Boulder, Colorado (NOAA and the Institute of Arctic and Alpine Research (INSTAAR)) and London (RHUL). Mole fraction measurements are traceable to the WMO X2004A scale for both NOAA and RHUL (see Dlugokencky et al. 2005). Intercomparison of isotopic measurement (δ<sup>13</sup>C<sub>CH4</sub>) between the INSTAAR and RHUL laboratories is by exchange of high pressure cylinders of air as detailed in Umezawa et al. (2017) and in Nisbet et al. (2016). δ<sup>13</sup>C<sub>CH4</sub> repeatability is ~±0.06‰ (INSTAAR) and ±0.05‰ (RHUL).

## 3. Results

Atmospheric methane grew very strongly in 2014-2017 (Figure 1 upper panel). Methane’s growth in 2014 was a remarkable 12.7±0.5 ppb, and, unusual for the record, the distribution of annual growth was global (Figures 2 and 3). Further very strong increases in 2015 (10.1±0.7ppb), 2016 (7.0±0.7 ppb) and 2017 (7.7±0.7 ppb) added to the methane burden already measured in 2014 (values based on NOAA CH<sub>4</sub> Trends: [www.esrl.noaa.gov/gmd/ccgg/trends\\_ch4/](http://www.esrl.noaa.gov/gmd/ccgg/trends_ch4/), accessed Oct 2018). Sustained annual growth of about 10ppb over the four year period was last observed in the 1980s, when the Soviet Union’s gas industry was developing very rapidly. The ~13 ppb increase in 2014 ranks among the very highest annual increments in the measured record.

Concurrently with the renewed growth since 2007,  $\delta^{13}\text{C}_{\text{CH}_4}$  has become significantly more negative by about 0.3‰ (Figure 1, lower panel). Overall globally, the bulk  $\delta^{13}\text{C}_{\text{CH}_4}$  ratio has shifted more negative by about -0.03‰/yr (see also Nisbet et al., 2016). This isotopic trend to more negative values, which has now been observed globally for more than a decade (Figure 1), reverses the sustained positive trend over the 19<sup>th</sup> and 20<sup>th</sup> Centuries (Francey et al., 1999).

The global nature of the changes is shown in Figure 2, from Alert (82°N) in the Canadian Arctic and Zeppelin (79°N), in Svalbard (Spitsbergen), both in the high Arctic; Ascension Island (8°S), just south of the equator in the remote central mid-Atlantic; and from the South Pole. Measurement data are shown without fitted curves to guide (and train) the eye, so that the reader can make an uninfluenced judgement on trends. Our measurements from all four stations are in agreement that the mole fraction of methane has risen, and that the  $\delta^{13}\text{C}_{\text{CH}_4}$  ratio has fallen significantly since 2007. However, since 2014, there is significant latitudinal variation in mole fraction and  $\delta^{13}\text{C}_{\text{CH}_4}$  trends (Figures 2 and 3), indicating a more complex regional picture. Most recently, over 2014-2017, in parallel with growth in the mole fraction of methane, the isotopic shift has continued in high northern latitudes, and also in Antarctica (see Figure 2, right panels). At Ascension Island, however,  $\delta^{13}\text{C}_{\text{CH}_4}$  fell in 2012 and in 2014, whereas in 2015 and 2016 the values stabilised (Figure 2c), with possible signs of resumed decline in 2017.

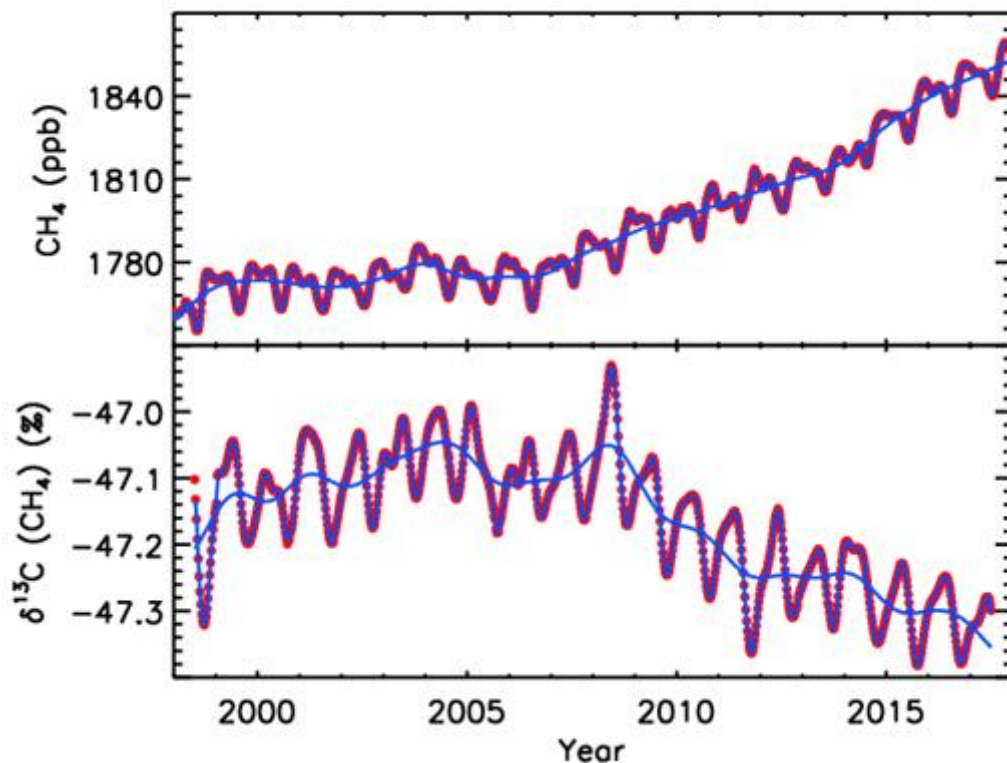
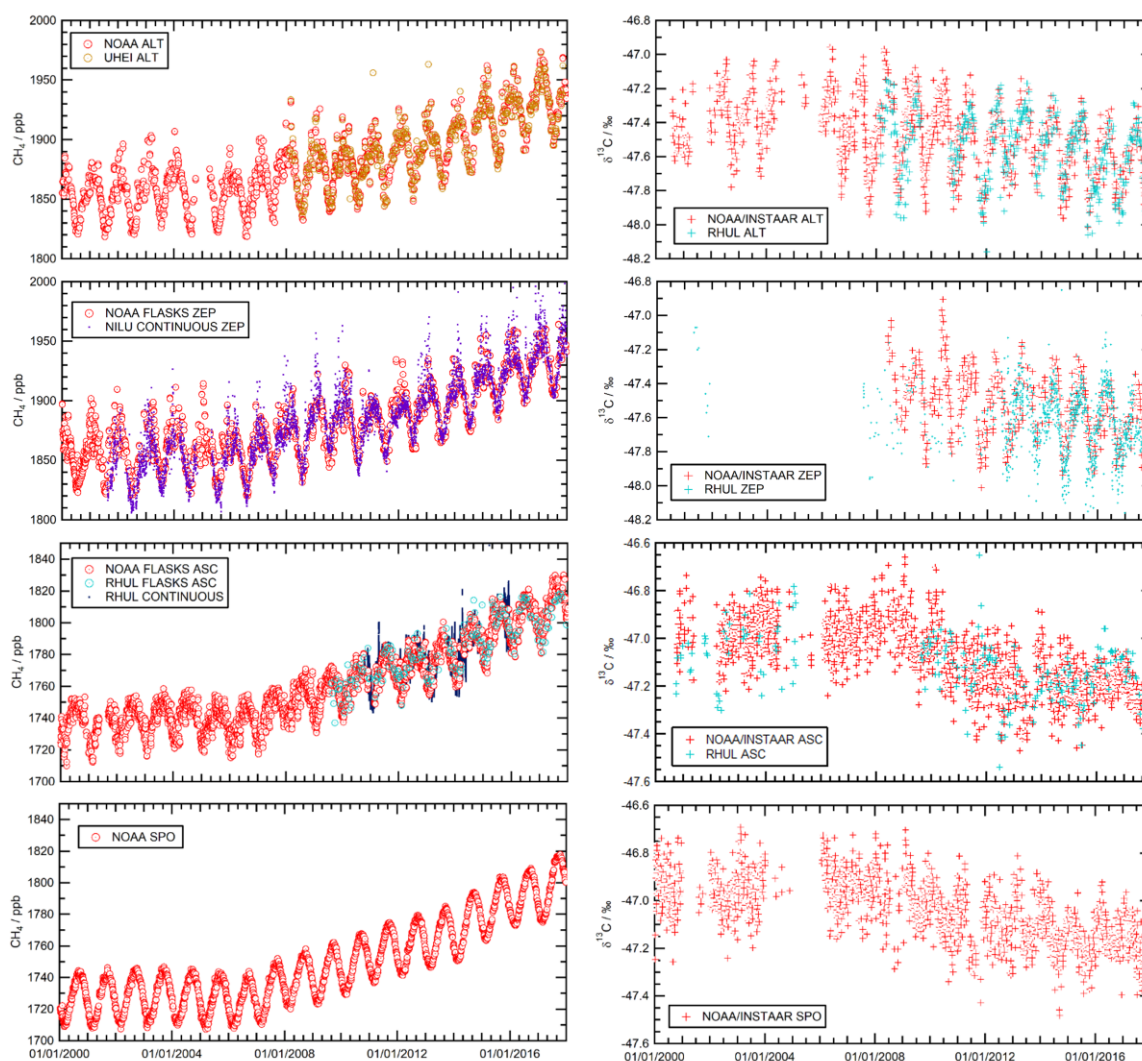


Figure 1. Atmospheric methane at Earth's surface in the remote marine troposphere (see Dlugokencky et al. (1994) for methods). Upper panel shows globally averaged surface atmospheric  $\text{CH}_4$  at weekly resolution (red and blue) and deseasonalized trend (blue), 2000-2017. Lower panel as above, but for globally averaged surface atmospheric  $\delta^{13}\text{C}_{\text{CH}_4}$ . Data from NOAA (see Section 2). X-axis tick marks denote January 1<sup>st</sup> of the year indicated.





**Figure 2.** Methane from Arctic, equatorial, and Antarctic sites. Left panels: mole fraction. Right panels:  $\delta^{13}\text{C}_{\text{CH}_4}$ .

a) Top: Methane at Alert, Canada (ALT; 82°N). b) Upper Middle: Methane at Zeppelin, Svalbard. (ZEP; 79°N). c) Lower Middle: Methane at Ascension (ASC; 8°S). d) Bottom: Methane at the South Pole (SPO).

X-axis tick marks denote January 1<sup>st</sup> of the year indicated. Note that both NOAA and RHUL results are similar, so trends are not single-lab artifacts. Measurements from NOAA/INSTAAR and RHUL, with additional flask samples from NILU (Zeppelin), and UHEI Institut für Umweltphysik (Alert) and RHUL/UK Met Office (Ascension). Note that y-axis ranges differ in different panels. Data are shown without guiding fits, so the reader can judge trends unguided.

#### 4. Extreme growth in 2014, 2015, 2016 and 2017: latitudinal analysis by zones

During the four year period 2014-2017, methane growth was sustained, each year bringing a major increment upon the remarkable growth of the previous year. Figure 3 shows contours of zonally-averaged  $\text{CH}_4$  growth rate ( $\text{ppb yr}^{-1}$ ) in the surface troposphere vs. sine of latitude and time. The sine-latitude plot is proportional to surface area and thus represents the global significance of each latitude zone. Methane mixes across the planetary troposphere in about a year and its lifetime is less than a decade (Lelieveld et al., 1998; Dlugokencky et al., 2011;

Dalsøren et al., 2016). In the late 20<sup>th</sup> and early 21<sup>st</sup> century pattern (Nisbet et al., 2016; Dlugokencky et al., 2011, 1994), single years of strong regional methane growth were typically followed by declines to the global background, as methane mixed from the surface, where it is measured, throughout the troposphere. In marked contrast, recent year-to-year growth has been so strong that growth has been recorded everywhere. Thus in the post-2000 context, 2014-2017 is very unusual, with few pronounced short-lived zonally averaged declines (Figure 3: for methodology see SI).

One hypothesis is that this growth has been driven by increased emissions. Regional pulses of strong seasonal emissions, whether from seasonal wetlands or during periods of high winter production from gas extraction, will disperse meteorologically in succeeding periods. As transport distributes emission pulses across the globe, strong growth years observed in specific latitude zones will likely be followed by short episodes of decline as regional emissions wax and wane, with atmospheric transport mixing the methane globally, including vertically, away from the surface where most observations are made.

Alternatively, if the growth in the mole fraction of methane has been driven by a decline in sinks, then the impact would be expected to be most marked in the tropical mid-troposphere, where methane destruction is most intense (e.g. see Table 1 below). The effects of reduced destruction would thus be likely to be most immediately observable in the tropics, but note that mid-troposphere air can travel far before interacting with the remote sites of the observation network, which mostly sample the marine boundary layer.

Figure 3a shows the methane growth rate from 1992-2018, from NOAA data. The figure includes, from 1992-2000, the later years of the long period of sustained growth that was observed from the early 1980s (when the global measurement network began) through to the end of the century. Then came the period of stability, growth in some years, decline in others, from the millennium until 2007. Strong growth resumed in 2007, becoming stronger from 2014. Figure 3b is the same as Fig. 3, but in detail showing the shorter time period 2005-2018. During the later years of the period of stability that lasted from the millennium, growth occurred in some years, decline in others. Strong growth resumed in 2007. Then, in the years from 2014, the focus of this report, the figure clearly shows the sharp increase in the growth rate.

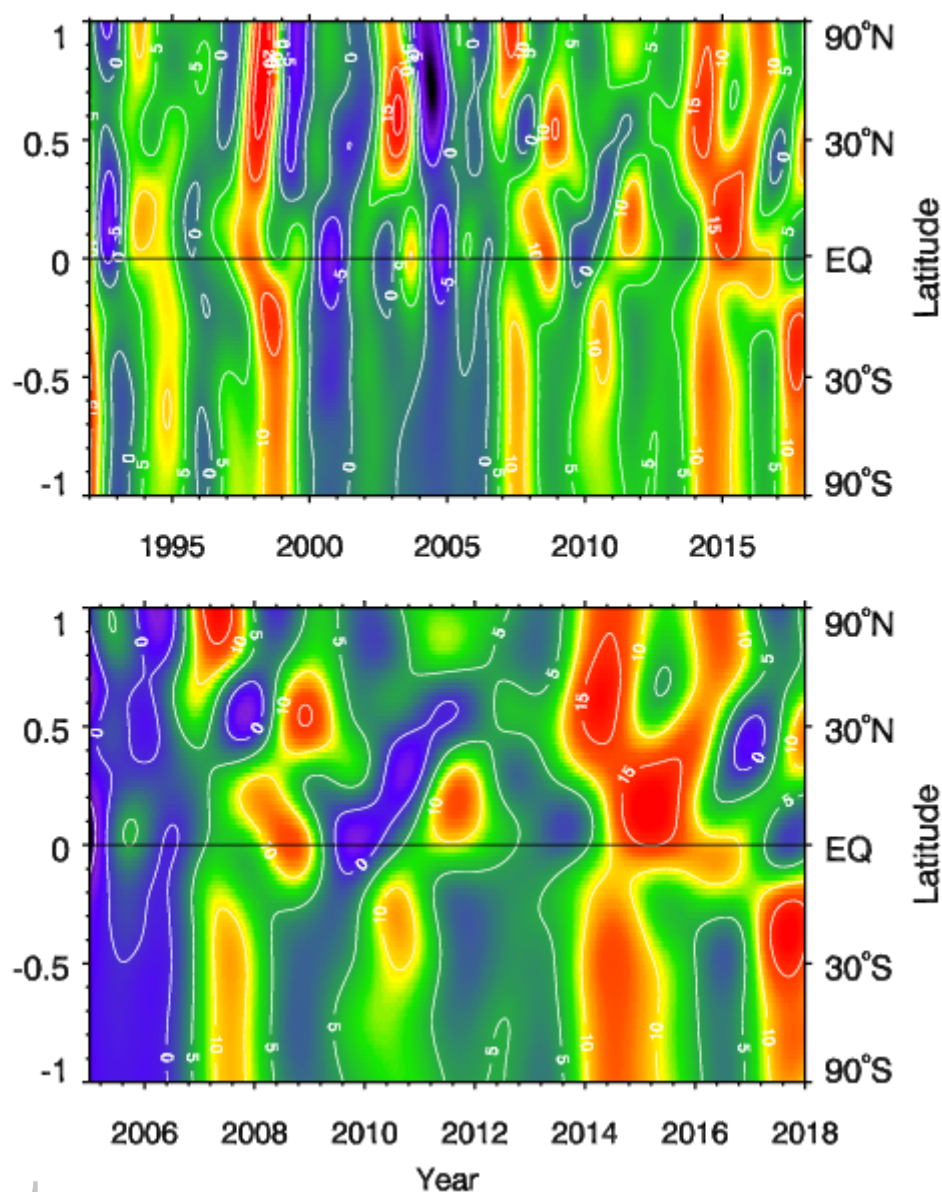


Figure 3. Zonally-averaged  $\text{CH}_4$  growth rate vs sine-of-latitude (equal area) and time:  
Fig. 3a 1992-2018.

Fig. 3b shows detail for 2005-2018.

Contours of growth rate (white lines) are shown every  $5 \text{ ppb yr}^{-1}$ . Warm colors show growth; cool colors decline. In this equal area plot note the importance of the zonal band under the sweep of the Inter-Tropical Convergence Zone (very roughly sin lat. 0.45 to -0.4). The Arctic and Boreal zones are above 0.8 sin lat. Note also that the contouring, which is based on deseasonalized trends, does not faithfully reproduce the timing of changes, and there are important end-effects. Thus the figure should be seen as illustrative, not definitive, especially in the past year. Data from NOAA.



#### ***4.1 Methane in the Arctic and Boreal zone (north of ~50°N).***

The mole fraction of methane grew strongly in the zone north of 50°N in the four year period between 2014-2017. In 2014, zonal growth as measured at NOAA's network of sampling sites (Dlugokencky et al., 2017; White et al., 2017) was high (>13 ppb) but the focus of higher growth was further south, around 40°N. Growth in 2015 was led by the northern tropics and the Arctic/Boreal zone lagged the rest of the world, but in 2016 the zone north of 50°N had the highest growth rate (>12 ppb) globally (Figure 3). Then in 2017, growth was focussed in the Southern Hemisphere and the Arctic/Boreal zone again lagged. The Alert and Zeppelin  $\delta^{13}\text{C}_{\text{CH}_4}$  records in 2014-2016 (Figure 2) show a continuation of the post-2007 shift to more negative  $\delta^{13}\text{C}_{\text{CH}_4}$  values (reported by Nisbet et al., 2016). In the entire post 2007 period, overall methane growth, measured at Environment Canada's Alert station and at NILU's Zeppelin station, was strong in the Arctic zone in mid-to-late 2007, but then growth in the latitudinal zone north of 50°N, though steady, tended to lag the rest of the world from 2008 until 2014, when growth was strong throughout the extratropical Northern Hemisphere, and again in 2016, but not in 2017 (Figure 3). Unfortunately  $^{14}\text{CH}_4$  and  $\text{CH}_3\text{D}$  monitoring are too limited to build a global picture.

Regional high-latitude emissions are primarily from wetlands in summer and gasfields in winter (Berchet et al. 2016; Fisher et al., 2011, 2017; France et al. 2016), though cold season wetland emissions may be more important than previously thought (Zona et al., 2016). There is no strong evidence for very large emissions from hydrates (Berchet et al., 2016; Lund Myhre et al., 2016). It should be noted also that at high latitudes the OH methane sink is relatively weak (see Table 1 below) and atmospheric transport is rapid and largely horizontal (Bousquet et al., 2011), thus changes in emissions have a more pronounced short-term impact on the zonal burden than at lower latitudes. As much of this region is warming disproportionately in global climate warming (Pitham and Mauritsen, 2014), Arctic and Boreal zone emissions would be expected to be responding sharply to increased warmth, but in the dryness and warmth, microbial soil methanotrophy would also be favoured. Nearly three decades of measurement at Barrow, Alaska, have shown little change in seasonal methane enhancements from the land sector, despite a significant increase in annual mean temperatures (Sweeney et al., 2016).

#### ***4.2 Methane in the populated north (30°N to ~50°N)***

Mid-northern latitudes led the strong global rise in 2014, when zonal methane growth was above 14 ppb. Methane growth was again 10 ppb in 2016. More recently, growth in this zone has been subdued. The isotopic trend to more negative  $\delta^{13}\text{C}_{\text{CH}_4}$ , began around 2005, and continued through the study period (see for example records from Mace Head, Ireland (53°N) and Terceira, Azores (39°N), in the NOAA data set <https://www.esrl.noaa.gov/gmd/dv/iadv/>), suggesting that growth was not led by fossil fuel emissions (which have relatively enriched  $\delta^{13}\text{C}_{\text{CH}_4}$ ) even though gas and coal extraction is very significant in this latitudinal zone (Schwietzke et al., 2016; Sherwood et al., 2017).

Overall this century, despite the increase in fossil fuel extraction and use, methane growth in this latitudinal zone has varied, with episodes of decline in 2004, 2006 and 2008. Apart from the notable growth in 2014 and 2016, the other episode of growth was in 2003, during a summer of extreme heat in Europe and major fires in Siberia (Trigo et al., 2005).

#### ***4.3 Methane in the Tropics and sub-Tropics (30°N to 30°S)***

Very strong methane growth (>10 ppb) took place in the tropics in 2014 and 2015; since then the tropical zonal annual growth rates remain high (around 8 ppb/yr). In 2014, growth in the southern tropics was nearly 13 ppb, and in 2017 the growth in the southern tropics was

extreme ( $>15$  ppb). These changes to the zonal methane burden took place during the remarkable episode of year-on-year temperature warming increments that were sustained through 2014, 2015 and 2016 (WMO 2016; UK Met Office 2018). Overall, growth since 2007 has been sustained in both northern and southern tropics, with prior episodes of strong increase in 2007-2008 (nearly 10 ppb), and 2010 (9 ppb), before the 2014-2017 period. Thus the tropics have played a major role in leading methane growth since 2007 (Figure 3), including during the 2015 *El Niño*.

As in the high north, the tropical  $\delta^{13}\text{C}_{\text{CH}_4}$  record at Ascension Island in the equatorial remote central Atlantic) shows a significant shift to more negative values (Figure 2). Here the  $\delta^{13}\text{C}_{\text{CH}_4}$  shift, which began at the end of 2007, may have temporarily halted around 2013 to the end of 2016. Since the start of 2017, the trend appears to be shifting again to more negative values, but future years of data will be needed to confirm this resumption of the trend to more depleted values. It should be noted that there is a significant worldwide tropical sampling deficiency for  $\text{XCH}_4$  and  $\delta^{13}\text{C}_{\text{CH}_4}$  measurement: for isotopes there is a critical lack of sustained multi-year tropical observational time series, which is arguably the weakest link in our understanding of global atmospheric methane. The only remote equatorial time series is from Ascension Island, which has almost invariant SE Trade winds from the deep South Atlantic, that later become the background input to Amazonian air masses.

#### **4.4 Methane in the South ( $30^\circ\text{S}$ to $90^\circ\text{S}$ )**

In parallel with global trends, growth in this zone was  $>12$  ppb in 2014, and 9 ppb in 2015. In 2016, growth dropped below 4 ppb, rising again in 2017 ( $>12$  ppb). This zone has proportionately much less methane-producing land mass than other zones, but is significant for the impact of the CI sink on the isotopic budget. As in the rest of the world, a sustained negative  $\delta^{13}\text{C}_{\text{CH}_4}$  shift is also observed at the South Pole (Figure 2). Unlike the Arctic stations, South Pole is far from major sources (see Table 1 below). Thus the South Pole time series portrays broad global trends: in contrast to the more variable Arctic record, the South Pole displays a smooth sustained trend to more negative  $\delta^{13}\text{C}_{\text{CH}_4}$  values throughout the period from 2007-2017, with the trend to more negative  $\delta^{13}\text{C}_{\text{CH}_4}$  becoming less pronounced since 2014.

### **5. Methane Sources and Sinks: The insight from carbon isotopes**

The various hypotheses to explain the current rise in methane are outlined in the introduction: the causes driving the current strong rise in methane are not fully understood, and may include changes in both sources and sinks (Bousquet et al., 2011; Dalsøren et al., 2016; Schaefer et al., 2016; Nisbet et al., 2016; Schwietzke et al., 2016; Saunio et al., 2017; Turner et al., 2017; Rigby et al., 2017; Worden et al., 2017). Although at present insufficient, the information required to constrain the relative causes of methane's observed rise lies in the geographic and seasonal spreads in both mole fraction data for methane and in corresponding isotopic measurements (Figures 2 and 3).

The continuing negative trend in  $\delta^{13}\text{C}_{\text{CH}_4}$  that began in 2007 is particularly elucidating as it implies a profound change in the evolving global budget, contrasting sharply with the two centuries of sustained positive trend in  $\delta^{13}\text{C}_{\text{CH}_4}$ . This earlier isotopic trend was sustained during methane's rapid 19<sup>th</sup> and 20<sup>th</sup> Century growth, contemporary with complex anthropogenic changes, that included fossil fuel emissions from natural gas, oil and coal extraction, perhaps increased biomass burning with increasing tropical agricultural populations (Crutzen and Andreae, 1990), and wetland drainage, in parallel with increasing ruminant populations.

Background atmospheric  $\delta^{13}\text{C}_{\text{CH}_4}$  is currently around -47‰. Methane removal by atmospheric hydroxyl, OH imposes a kinetic isotope effect of around 3.9‰ (Saueressig et al., 2001) or 5.4‰ (Cantrell et al., 1990), which implies a bulk global emission  $\delta^{13}\text{C}_{\text{CH}_4}$  value from all sources contributing to the current burden of between -54‰ and -52‰ (Nisbet et al., 2016). However, interpretation of isotopic observations is complex, because of the multi-decadal timescale of isotopic equilibration (Tans, 1997). In the short-term, emissions with  $\delta^{13}\text{C}_{\text{CH}_4}$  between -47‰ and -53‰ initially help move the total atmospheric burden negative, but as the emissions are destroyed by OH, on a decadal time-scale, the long-term effect is to move the bulk global burden positive. Thus the sustained post-2007 negative isotopic shift, if driven primarily by an emission increase and over a long period, a decade or more, implies increasing emissions from sources more negative than about -53‰.

Methane *sources* vary with latitudinal zone. Isotopic signatures of sources are summarised by Dlugokencky et al. (2011) and in more detail by Sherwood et al. (2017), Zazzeri et al., 2017, and Brownlow et al. (2016). Much depends on vegetation type: temperate vegetation is mainly of C3 plants, which are very selective for  $^{12}\text{C}$ . Tropical biomes, especially the savannas and wetlands, typically include abundant C4 grasses (e.g. papyrus). These evolved during low  $\text{CO}_2$  periods and their carbon capture from the air is less selective against  $^{13}\text{C}$ : thus their biomass has higher  $^{13}\text{C}$  contents. Globally, our knowledge of  $\delta^{13}\text{C}_{\text{CH}_4}$  signatures of methane released into open air is very inadequate (Feinberg et al., 2018).

Emissions from the Arctic and boreal zone north of  $50^\circ\text{N}$  come dominantly from widespread summer C3 wetlands ( $\delta^{13}\text{C}_{\text{CH}_4}$  about  $-70\pm 5\text{‰}$ ) and also from releases, especially in winter, from the very large Arctic gas fields ( $\delta^{13}\text{C}_{\text{CH}_4}$  about  $-50\pm 5\text{‰}$ ) (Fisher et al., 2011; 2017; France et al., 2016). This region also hosts permafrost and gas hydrates, though emissions from gas hydrates and seeps currently seem to be lower than previously estimated (Lund Myhre et al., 2016; Berchet et al., 2016; Thornton et al., 2016a). Double-counting in emissions inventories may mean that Arctic high latitude emissions may have been overstated, and may currently be as low as roughly 4% of the global total (Thornton et al., 2016b), although regional inverse modeling suggests up to 15% of global methane emissions, both natural and anthropogenic (including the giant Siberian Arctic gas fields), come from the region north of  $50^\circ\text{N}$  (Thompson et al., 2017). The question of the region's significance on a global scale thus remains open.

Methane emissions in the zone from  $50^\circ\text{N}$  to the Tropic of Cancer include both strong biogenic and anthropogenic emissions. These come from: extensive Russian and Canadian C3 wetlands ( $\delta^{13}\text{C}_{\text{CH}_4} \sim -65\pm 5\text{‰}$ ); from the natural gas industry ( $\delta^{13}\text{C}_{\text{CH}_4} \sim -50$  to  $-30\text{‰}$ ), with much of the world's production in the USA, Russia and the Middle East; and from coal production ( $\delta^{13}\text{C}_{\text{CH}_4} \sim -40\pm 15\text{‰}$ ), especially in China.  $\delta^{13}\text{C}_{\text{CH}_4}$  values given here are rough overall estimates based on Schwietzke et al. (2017), Brownlow et al. (2017), Zazzeri et al. (2016) and Ganesan et al. (2018).

China may have been responsible for up to 40% of global methane growth between 2003-2005 and 2007-2010 (Thompson et al., 2015), and was likely an important contributor to growth in the post-2014 years. In particular, using an atmospheric Bayesian inversion, Thompson et al. (2015) inferred that total emissions from all sources in East Asia grew from  $43\pm 4$  Tg in the year 2000 to  $59\pm 4$  Tg in 2011 much of which was because China's annual emissions grew from  $39\pm 4$  Tg to  $54\pm 4$  Tg in the period. However, Feinberg et al. (2018) show it is hard to reconcile inventory estimates of growth in Chinese coal production (e.g. EDGAR 2017) with the observed isotopic trends, unless China is switching to lower quality coals with more negative  $\delta^{13}\text{C}_{\text{CH}_4}$ .

In the South American and African tropics, globally significant wetland emissions occur

(Melton et al., 2013). These are primarily from the Amazon and Pantanal wetlands, and in Africa from the Congo, Nile, Zambesian, and other C4 papyrus wetlands: their  $\delta^{13}\text{C}_{\text{CH}_4}$  signatures are poorly studied but likely  $\sim -55\pm 10\text{‰}$  (Brownlow et al., 2017). Emissions from tropical wetlands vary as wetland area responds to changes in precipitation, and increase exponentially with temperature, provided water and nutrients are available (Gedney et al., 2004; McNorton et al., 2016a; Westerman & Ahring, 1987). The  $\delta^{13}\text{C}_{\text{CH}_4}$  values of cattle breath also contribute to tropical methane emissions. These values are poorly studied, and will depend on proportions of C4 fodder such as maize and sugar cane tops, but are likely  $\sim -55\pm 10\text{‰}$  (e.g. Brownlow et al., 2017). Tropical cattle prosper in wet seasons.

Similarly, tropical Asia has abundant cattle (especially in India) and wetlands (Ganges Delta and S.E. Asia including Borneo), as well as extensive gas industries, and large-scale coal mining in India and Indonesia. It is possible that, as pointed out by Worden et al. (2017), a reduction of biomass burning emissions of a few Tg/yr may have occurred and masked increasing fossil fuel-related emissions in the global emissions budget. This is discussed further below (Sect. 8.2).

In the extra-tropical Southern Hemisphere, natural emissions are from cattle, biomass burning, and some wetlands, with significant industrial emissions from Australia (gas, coal) (Zazzeri et al., 2016) and South Africa (coal).

Geological emissions also occur worldwide, but may be much smaller than hitherto estimated (Petrenko et al., 2017).

The main sink of atmospheric methane, destruction by atmospheric hydroxyl (OH), occurs especially in the bright moist tropical mid-troposphere, where OH, which has a very short lifetime ( $\sim 1$  s), is most active. OH is thought to be well buffered, not sensitive to perturbations by natural or anthropogenic emission changes. OH inter-annual variability can have important impact on methane growth rates (e.g. see McNorton et al. 2016b). As OH destruction is likely modulated by strong changes in meteorology (Holmes et al., 2013), such as may have taken place in the past decade in tropical mid-tropospheric moisture and cloud structure, it is possible that significant inter-annual variability in OH destruction may indeed occur, though detailed mechanisms capable of driving such large changes in the oxidative capacity of the atmosphere have not been proposed. However, in this context, Nicely et al. (2019) and Lelieveld et al. (2016) agreed with Montzka et al., (2011) in finding that interannual variability of OH has probably been fairly small.

The uncertainty in our understanding of OH is illustrated by the findings of Naus et al., (2019), who note a number of biases caused by using a box model, resulting from a combination of variations in transport and variations in source/sink distributions; when those biases are corrected using a 3D model, they find an opposite trend in OH. In their work, the sensitivity of interannual OH anomalies to the biases is modest (1–2%), relative to the significant uncertainties on derived OH (5–8%). In an inversion implementing all bias corrections simultaneously, they did find a shift to a positive OH trend over the 1994–2015 period.

Reduced destruction by lesser sinks may also have contributed to the rise in the global methane burden (Nisbet et al., 2016). These smaller sinks include soil methanotrophy (Holmes et al., 1999; Serrano-Silva et al., 2014), caves in karst limestone topography (Mattey et al., 2013), and Cl from marine aerosols, especially in the Southern Hemisphere with its large oceans (Hossaini et al., 2016). Soil moisture and hence methanotrophy vary with rainfall, temperature, and the height of a water table, so that drought can increase rates of consumption of atmospheric methane by a factor of  $>4$  (Davidson et al., 2004). The Cl sink



(Allan et al., 2007; Hossaini et al., 2016) is likely affected by varying wind conditions in the marine boundary layer.

## 6. Testing the hypotheses that may explain the recent rise in methane

### 6.1 *The hypothesis that biogenic methane emissions (mainly from wetlands and cattle) have increased, while destruction by OH has been relatively constant*

The first hypothesis to explain the methane growth and negative  $\delta^{13}\text{C}_{\text{CH}_4}$  trend is that much of the 2014-2016 growth could have come from increased biogenic emissions, while destruction rates remained steady. Recent years have been very warm and increased emissions would be expected from widespread tropical wetlands, supplemented by expanding populations of tropical ruminants, well-fed on good vegetation growth (Nisbet et al., 2016; Schaefer et al., 2016), and perhaps the rapid expansion of landfills, many unregulated, in the tropics. For example, Thompson et al. (2018) inferred that between 2006-2014 microbial sources increased by  $35 \pm 12$  Tg/yr, and fossil fuel sources by  $13 \pm 8$  Tg/yr, while biomass burning decreased by  $3 \pm 2$  Tg/yr and soil oxidation increased by  $4 \pm 6$  Tg/yr. If the hypothesis that growth is primarily driven by increasing emissions from biogenic sources is correct, then the continuing negative shift in 2014-2016 suggests further enhancement of the biogenic inputs.

The hypothesis that wetlands have led the recent growth is consistent with evidence for a post-2006 increase in wetland emissions (+3%), mainly from the Tropics (McNorton et al., 2016a; Zimmermann et al., 2018; Parker et al., 2018). Moreover, there is substantial evidence for the expansion of the meteorological tropics, especially in summer (e.g. Davis & Birner, 2017, Allen et al., 2014). Ganesan et al. (2018) showed the need for better isotopic knowledge of tropical wetland emissions before the large uncertainties in fluxes derived from inversion models can be reduced. Using a spatially resolved source signature map, they found mean  $\delta^{13}\text{C}_{\text{CH}_4}$  of northern wetland methane emissions to be -67.8‰, and mean tropical wetland emissions to have  $\delta^{13}\text{C}_{\text{CH}_4}$  -56.7‰.

The years 2014, 2015 and 2016 were of exceptional warmth, each year surpassing the previous as the warmest on record (WMO 2016; Rahmstorf et al., 2017). The extreme 2014 methane growth event took place in a year that was 0.99°C warmer compared to 1880-1909 conditions (NCDC, 2017). In 2015 (1.13°C warmer than 1880-1909 conditions) the extreme warming trend continued, and again in 2016 (1.25°C warmer than 1880-1909 conditions), a third year of record warmth, each year having surpassed the prior year, successively setting global temperature records. Then 2017, which was 1.16°C above 1880-1909 norms, was the warmest non-*El Niño* year on record (UK Met Office 2018). The coincidence of methane growth with these extreme years raises the question of the extent to which changes in temperature and precipitation may be contributing to the sustained growth rate in methane. This question is important because this type of climate system feedback effect is not explicitly parameterised in coupled atmosphere-ocean general circulation models. The temperature dependence of microbial methane emissions (McNorton et al., 2016a; Westerman & Ahring, 1987; Gedney et al., 2004) suggests growth came as a response to that warmth. The very strong methane growth in 2014 may thus reflect the very warm temperatures compared to previous years.

Tropical wetland emissions respond quickly to warmth and increased precipitation, and are relatively depleted in  $^{13}\text{C}$  compared to the atmosphere (Nisbet et al., 2016; Dlugokencky et al., 2011; Brownlow et al., 2017). Flooding can be significant in expanding the surface area and hence emissions of key tropical wetlands. Mass transfer of water from ocean to land was tracked by the Gravity Recovery and Climate Experiment (GRACE) satellite (Boening et al., 2011). Analyzing GRACE data, Parker et al. (2018) found large discrepancies between

estimates of CH<sub>4</sub> emissions based on GOSAT observations, and also emission estimates from land surface models, which may have badly underestimated tropical wetland methane emissions, by tens of teragrams.

Flooding may partly explain global methane growth in 2014-8. Intuitively, warmer and wetter conditions would increase wetland emissions. Methane emission depends logarithmically on temperature and linearly on inundation areas. The very extensive wetlands in Bolivia had extreme flood events in early 2014 (Ovando et al., 2015), filling wetlands before the onset of intense heat in 2014. Growth in 2015 took place in the powerful 2015-2016 *El Niño*, one of the strongest on record (L'Heureux et al., 2017; WMO 2016). There were major rainfall anomalies, with increased rainfall in equatorial Africa and Angola/W.Zambia, as well as Paraguay, but drought in Amazonia (L'Heureux et al., 2017). Thus the wetland record is complex. However, both Amazonia and tropical Africa, and indeed most of the planet, were extremely warm (L'Heureux et al., 2017). Thus methane emissions may have grown as biogenic sources responded to both warmth and flooding. Similarly, 2016 was again a year of record warmth and the sustained methane rise in 2016, though not as strong as in the preceding two years, may also be primarily a biogenic response to the third successive increase in world-wide temperatures.

Ruminant emissions, especially cows, may also have played a major role (Schaefer et al., 2016, Wolf et al., 2017). Ruminant emissions are isotopically almost indistinguishable from wetland emissions. A cow is a walking wetland at 37°C: methane from C4 (papyrus) equatorial wetlands and methane from cows fed with C4 fodder (maize, sugar cane waste) are isotopically similar and only distinguishable by back trajectory location and population analysis. Thus, it is probable that both wetlands and cows are contributing to current growth in tropical emissions. However, it should be noted that in India, the nation with by far the world's highest cattle population, atmospheric observations suggest there has been little or no growth in methane emissions between 2010 and 2015 (Ganesan et al., 2017). In other nations with very large ruminant populations (Brazil, China, US, Ethiopia, Argentina, S. Sudan) this may not have been the case.

If the rise in methane and the simultaneous negative  $\delta^{13}\text{C}_{\text{CH}_4}$  shift have been driven by increased biogenic emissions, then it is feasible that an absolute increase in fossil fuel emissions may have coincided with a concurrent reduction in the fossil-fuel proportion in the total methane emission source mix (Schwietzke et al., 2016; Schaefer et al., 2016; Nisbet et al., 2016). Methane emitted during extraction, transport, and combustion of coal and natural gas has characteristically more positive  $\delta^{13}\text{C}_{\text{CH}_4}$  (say -35 to -50‰) than the bulk -53‰ global methane source (Dlugokencky et al., 2011; Zazzeri et al., 2016), although it should be noted that the global mean for fossil fuel-related emissions has recently been shown to be around -44±0.7‰, isotopically somewhat more negative than hitherto assumed (Schwietzke et al., 2017). Indeed, methane from some Arctic gas fields and also Australian coalfields has relatively negative  $\delta^{13}\text{C}_{\text{CH}_4}$  around -50‰ or less (Dlugokencky et al., 2011; Zazzeri et al., 2016), but these high latitude and southern emissions are geographically unlikely to have driven global methane growth, which has been led primarily from the wet tropics and low northern temperate latitudes.

The inference, that the share of global emissions coming from fossil fuels is dropping, sharply contradicts inventory evidence that global gas and coal production is growing (BP Statistical Review 2017, EDGAR 2017) and that methane emissions from fossil fuels have risen substantially this century (EDGAR 2017, USEPA 2012). However, there is evidence (Schwietzke et al., 2016) that natural gas emissions per unit of production have declined significantly in recent years, and rapid improvements and investment in leak detection and reduction have likely cut the percentage of gas leaked from gas industry production facilities.

Observation of ethane trends (Hausmann et al., 2016) suggests growing emissions of methane from the hydrocarbon industry. Helmig et al. (2016) reported a recent increase in ethane emissions and that methane/ethane ratios could suggest that Northern Hemisphere oil and gas industry methane emissions had increased significantly recently (mid-2009 to mid-2014). But this assumes that ethane/methane ratios are constant and that they are coupled at large spatial scales: nevertheless, both assumptions can be challenged. Helmig et al. (2016) pointed out that such a postulated increase in gas emissions is inconsistent with observed leak rates and the isotopic results such as those cited above. Note also that ethane is released with methane in coal mining; thus even if leaks have been reduced from natural gas facilities, the ethane record may reflect a growing contribution from coal mining, which has expanded greatly in East and South Asia in recent years.

Thus although fossil fuel emissions of methane are very large and a more major part of the global budget than previously thought (Schwietzke et al., 2016), a simple explanation of the observed recent shift in  $\delta^{13}\text{C}_{\text{CH}_4}$  to values more depleted in  $^{13}\text{C}$ , assuming the OH sink has been relatively unchanged, is that fossil emissions are falling as a proportion of the total methane emission (Schaefer et al., 2016; Nisbet et al., 2016), although they may have increased in absolute terms.

## ***6.2 Has either a shift in fossil fuel $\delta^{13}\text{C}_{\text{CH}_4}$ or a decline in biomass burning masked an overall increase in fossil fuel emissions, while sinks remained relatively constant?***

It is possible overall that methane emissions from fossil fuel sources have increased, but that the isotopic impact of this increase has been ‘masked’ either because fossil fuel sources have overall become isotopically more negative, or because there has been a parallel decline in isotopically more positive output from biomass burning, or via some combination of both factors. This hypothesis assumes sinks have been steady.

Has the bulk  $\delta^{13}\text{C}_{\text{CH}_4}$  of global fossil fuel sources become more negative in the past decade? It is very difficult to comment on this, as measurements are inadequate. But the marked global trend away from coal and towards natural gas might support the notion. Methane from coal mines is isotopically variable: Zazzeri et al. (2017) suggested  $\delta^{13}\text{C}_{\text{CH}_4}$  values for bituminous coals of -65‰ for open cast mines and -55‰ for deep mines, while methane from anthracite mines ranged from -40‰ to -30‰. Thus growth in open cast mines and closure of deep mines would shift the bulk  $\delta^{13}\text{C}_{\text{CH}_4}$  of emissions to more negative values. However,  $\delta^{13}\text{C}_{\text{CH}_4}$  shifts driven by changing methane fluxes from coal mines may have been counteracted by the rapid growth in extraction of natural gas. Schwietzke et al. (2016) have discussed these questions in detail.

A parallel contributory hypothesis is that a decline in biomass burning emissions has masked a strong increase in fossil fuel emissions, given evidence suggesting reduced burning in the 2001-2014 period (Worden et al., 2017). As noted above, Thompson et al. (2018) also infer that fossil fuel sources increased in 2006-2014, while biomass burning decreased. Methane from fires, especially in tropical savannas dominated by C4 grasses, is much enriched in  $\delta^{13}\text{C}_{\text{CH}_4}$  (Dlugokencky et al., 2011; Brownlow et al., 2017); thus pulses of methane from tropical savanna and peatland fires ( $\delta^{13}\text{C}_{\text{CH}_4}$  around -30‰ to -15‰), have strong leverage on ambient values. If so, declines of emissions from smaller but more heavily  $\delta^{13}\text{C}_{\text{CH}_4}$  positive sources such as biomass burning (especially of tropical C4 grasses) and deep coal-mining could have ‘masked’ the isotopic impact of strong growth in emissions from industrial natural gas extraction, transmission, and use. Thus a strong increase in emissions from natural gas leaks could indeed have occurred, as would be inferred from production growth (BP 2017), but with the isotopic impact hidden by the parallel decline in methane from fires. However, in this context it should be noted that there is evidence for doubling of ammonia emissions from

biomass burning in the 2015 *El Niño* in S.E. Asia (Whitburn et al., 2016). Possibly the apparent standstill in the tropical  $\delta^{13}\text{C}_{\text{CH}_4}$  trend in the four years 2014-2017 (see section 4.3) may reflect increased biomass burning methane emissions in this period.

What is clear is that because the  $\delta^{13}\text{C}_{\text{CH}_4}$  signature of biomass burning is relatively so positive, the global negative isotopic shift excludes an *increase* in biomass burning as an explanation for the methane rise, despite the warmth of 2015-2017, and this is supported by direct measurements of fire extent (van der Werf et al., 2017). Burnt area is not necessarily an accurate guide to methane emission from fires: the volume of the fuel load is arguably more important, especially in African grasslands. Carbon monoxide is another proxy for biomass burning, and has been broadly stable or declining at Ascension (NOAA data) and Zeppelin (NILU data). Much of this CO decline is in part as a result of controls on vehicle emissions (Yin et al. 2015), but a decline in biomass burning emissions is nevertheless plausible. Although Lelieveld et al. (2016) and Nicely et al. (2018) concluded that global OH is insensitive to perturbations by natural or anthropogenic emission changes, falling CO may also affect ratio of CO to NO<sub>x</sub> regionally, and hence OH and the lifetime of methane in that region (see below).

Currently the data are inadequate to reach firm conclusions on these hypotheses. The more general inference is that the observed negative shift in  $\delta^{13}\text{C}_{\text{CH}_4}$  is not necessarily incompatible with an increase in overall fossil fuel emissions, if this increase has coincided with a shift from coal to gas, and/or a synchronous strong increase in emissions from sources with more negative  $\delta^{13}\text{C}_{\text{CH}_4}$  such as wetlands, and/or a decrease in a more positive  $\delta^{13}\text{C}_{\text{CH}_4}$  emissions, such as from biomass burning, or all factors together.

### **6.3 Possible methane sink strength and lifetime changes driven by a decline in the oxidative capacity of the atmosphere.**

An alternative explanation for the methane rise is that sinks have declined and thus methane's lifetime has increased, so that a given source supports a larger burden in the atmosphere. A decline in oxidation would lead to a negative  $\delta^{13}\text{C}_{\text{CH}_4}$  shift of the global methane burden. This would mean that methane's lifetime has increased such that a steady emission flux supports a higher atmospheric burden. Note that the various hypotheses are non-exclusive: this could be a complementary hypothesis, concurrent with a change in sources.

Using a multi-species non-linear Bayesian two-box inversion model to fit a smoothed deseasonalized hemispheric methane record, Turner et al. (2017) found that the most likely explanation for renewed growth in atmospheric methane was a 25Tg/yr *decrease* in total annual global methane emissions from 2003 to 2016, which was offset by a 7% decrease in global mean hydroxyl (OH) concentrations. However, this contrasts with the findings of Montzka et al. (2011) and Lelieveld et al. (2016) that OH is buffered, and Dalsøren et al. (2016) who found increasing OH from 2003-2005, and stable OH during 2006-2012. As shown below, a decrease in the more highly fractionating removal by Cl could also explain much of the observed changes in methane mole fraction and  $\delta^{13}\text{C}_{\text{CH}_4}$ , but it has to be a very large relative change in Cl together with some change in OH as well.

Rigby et al. (2017) also used a Bayesian study to model  $\text{CH}_3\text{CCl}_3$  emissions as a proxy for OH, and explore the methane growth up to 2014. They found that the likeliest explanation of the growth was that OH abundance grew by around 10% between the late 1990s and mid-2000s, and then fell by a similar magnitude from the mid-2000s to 2014. This is consistent with results from an ensemble of biogeochemical models (Poulter et al., 2017) that suggested the methane rise up to 2012 involved *decreased* tropical emissions. This would support explanations of growth that involved a combination of increasing fossil fuel and agriculture-related methane emissions and a decrease in the atmospheric oxidative sink.



Both studies (Rigby et al., 2017; Turner et al., 2017) commented that the problem is under-determined: the long-term time series measurements of methane and  $\delta^{13}\text{C}_{\text{CH}_4}$  are inadequate, as is the knowledge of the regional variability of isotopic signatures of emissions from specific sources. Although not favoured by their interpretations, neither Turner et al. (2017) nor Rigby et al. (2017) could discard the null hypothesis that OH has remained steady in recent years. Neither suggested any mechanism for the postulated changes in OH, although Dalsøren et al. (2016) point to the impact on OH of variations in  $\text{NO}_x$ , especially noting large increases in OH over Southeast Asia, mainly due to strong growth in  $\text{NO}_x$  emissions; more generally they commented that an increase in  $\text{NO}_x$  increases global OH as long as it takes place outside highly polluted regions.

These studies by Turner et al. (2017) and Rigby et al. (2017) raise the important question of what is happening to OH, the major methane sink. There is evidence for relatively small inter-annual variation (Montzka et al., 2011), although in some years changes  $>4\%$  may have occurred. Yet the inversion models of Turner et al. (2017) and Rigby et al. (2017) suggest it fluctuates very much more. In contrast, Naus et al. (2019), using output from 3D model simulations, found only small differences up to 1.3%, averaged over 1994-2015, in OH anomalies relative to the full uncertainty envelope (5-8%). The question is important not only for understanding the methane budget but also because the implications of any postulated substantial declines in the oxidative capacity of the atmosphere are profoundly worrying – the troubling inference is that the air is losing some of its self-cleansing mechanism. Discussing this debate, Prather and Holmes (2017) pointed to the need to push the box modelling nearer to the real world, to try to understand how climate changes or human actions can alter methane emissions and OH trends.

In both modelling studies, the main indicator species used for OH is methyl chloroform,  $\text{CH}_3\text{CCl}_3$ , but there are major problems in using this proxy. Although manufacture of this Montreal-Protocol species is formally prohibited, it is an excellent industrial solvent (e.g. for typing correction fluid), and thus there may be emissions continuing from landfills, especially in developing countries (Talaiekhosani et al., 2016). As possible parallels, Montzka et al., (2018) report an unexpected and persistent ongoing increase in prohibited CFC-11 emission and Vollmer et al. (2018) come to similar conclusions for ongoing CFC-13 output, possibly from East Asia. Until ongoing  $\text{CH}_3\text{CCl}_3$  emissions are better quantified, or a new proxy for OH is found, our knowledge of OH and thus the oxidative capacity of the atmosphere, will remain very poorly constrained. In this context, NIWA-New Zealand's  $^{14}\text{CO}$  data that should be published shortly may provide a better proxy.

Dalsøren et al. (2016) investigated various factors influencing OH and found the ratio  $\text{NO}_x/\text{CO}$  is the most important. Until recently this ratio has increased, but new inventories from the Community Emission Data System (CEDS) (Hoesly et al., 2018) indicate reduction in  $\text{NO}_x$  emissions and stable CO emissions. This suggests a reduction in OH over the last few years, but data are inadequate to test the suggestion robustly. There are reasons to expect OH to be increasing in some places and decreasing in others (Lelieveld et al., 2004, 2016; Dalsøren et al., 2016), which could keep the global average constant with little effect on methane, but still have significantly different effects on short lived trace gases depending on the spatial distributions and seasonality of their sources.

In this context, Thompson et al. (2018) found that OH appears not to have contributed significantly to recent methane growth. Further insight into the OH problem came from Nicely et al. (2018) who showed that atmospheric OH is probably currently well buffered, with mean anomalies of 1.6%. Although OH would be expected to decrease as methane rises, this effect is countered by the tropical widening as the Hadley cells expand (Allen et al., 2014), combined with the influences of changing  $\text{H}_2\text{O}$ ,  $\text{NO}_x$  and overhead  $\text{O}_3$ .

The timing of the isotopic shift provides a separate test of the ‘declining OH sink’ hypothesis to explain the rise in methane. Methane destruction by OH changes the  $\delta^{13}\text{C}_{\text{CH}_4}$  value of the methane burden only slowly, over many decades (Tans, 1997), and inter-hemispheric  $\delta^{13}\text{C}_{\text{CH}_4}$  equilibration is slow. Methane’s large seasonal cycles in emissions and removal mean that  $\delta^{13}\text{C}_{\text{CH}_4}$  never reaches equilibrium. Moreover, the seasonal cycles of mole fraction and  $\delta^{13}\text{C}_{\text{CH}_4}$  may be different, and perturbation impacts on  $\delta^{13}\text{C}_{\text{CH}_4}$  are sensitive to the time of year.

In contrast to the slow timescale of isotopic adjustment to changes in OH destruction, the ‘emission-driven’ explanation of the methane rise, that pulses of emission inputs from biogenic sources took place, immediately changes regional and hemispheric  $\delta^{13}\text{C}_{\text{CH}_4}$ . Imagine a major burst of wetland methane at the end of a wet season: this would cause regional  $\delta^{13}\text{C}_{\text{CH}_4}$  to plummet. This is exactly the type of rapid shift observed in the isotopic records illustrated in Figure 2. Sharp drops and peaks in  $\delta^{13}\text{C}_{\text{CH}_4}$  are visible, superimposed on the seasonality. These shifts are visible in multiple sites and are thus not just caused by local one-off meteorological variability. Such shifts are recorded (e.g. in 2014) by the coherent drops shown by the whole cortège of hemispheric stations, even in extratropical stations such as South Pole (Figure 2), and Cape Point (South Africa) (see also Nisbet et al., 2016), which sample air masses far away from any local sources.

Given the slow timescale of  $\delta^{13}\text{C}_{\text{CH}_4}$  equilibration (Tans 1997) it is hard to escape the inference that these shifts are too rapid to be solely a response to OH variation through the long-term kinetic isotopic impacts of OH destruction. More likely, the very rapid isotopic shifts in the record are driven, at least in part, by regional or hemispherical-scale source inputs. By fitting smoothed and deseasonalized curves to the measurement data, especially when using 1-box or 2-box models, inverse models can find apparent OH-led solutions for smoothed very long-term trends, but the very act of data smoothing raises the question whether critical emission-related insights from shorter-term variation are being lost.

Many of the possible changes in methane sources and/or sinks are direct or indirect responses to meteorological change. Wetlands may be warming, tropical ruminants may be flourishing more under better rainfall, or OH may be responding to change in the tropical troposphere. The negative shift in  $\delta^{13}\text{C}_{\text{CH}_4}$  suggests climate-linked processes are the most plausible explanations of the isotopic shift. Such decadal-scale feedbacks, especially in the tropics may be secondary, not primary, anthropogenic forcings. The hypotheses that invoke primary anthropogenic inputs are the possibility that emissions from ruminants (primarily cattle) have increased, that a decline in biomass burning has masked the isotopic consequences of an increase in fossil fuel leaks, or that anthropogenic pollution may have caused OH decline.

## 7. Running budget model

To address the causes of methane growth, ‘running budget’ modelling of data from 1998–2018 was carried out, extending the methodology developed by MRM in Nisbet et al. (2016) and summarised here in the Supporting Information. The previous analysis (Nisbet et al. 2016) investigated whether observed changes in methane mole fraction and isotopic content, averaged over four semi-hemispheric regions, could be explained by either of two end-member hypotheses: ‘*increases in sources*’ or ‘*declines in sinks*’. Optimisation of fits to the observed seasonal cycles and trends in atmospheric data, by adjusting interannual changes in sources or sinks, has now been extended to use a significant revision of the highly isotopic fractionating removal of  $\text{CH}_4$  by Cl (Hossaini et al., 2016), a more detailed treatment of the soil sink (Curry, 2007), and to include the more recent atmospheric data. Our analysis of semi-hemispheric source budgets is deliberately using a coarser resolution than is done in inverse models.

As noted above, Turner et al. (2017) considered that the most likely explanation of recent methane growth was that OH had declined, and Rigby et al. (2017) also found it very likely that an OH decline had played a major role in methane growth, while Naus et al. (2019) found that OH variation was likely to have been less. In contrast to Turner et al.'s (2017) modelling of smoothed multi-year trends, the inclusion of seasonal cycles in the analysis makes a significant difference when  $\delta^{13}\text{C}_{\text{CH}_4}$  data are considered. Neither  $\text{CH}_4$  nor  $\delta^{13}\text{C}_{\text{CH}_4}$  are in equilibrium, and the two parameters have quite different response rates to budget changes (Tans 1997), being nonlinear for  $\delta^{13}\text{C}_{\text{CH}_4}$ . Thus smoothing the record removes important signals. Nisbet et al. (2016) showed that a decrease in OH could indeed explain a shift in trend from  $\delta^{13}\text{C}_{\text{CH}_4}$  growing more positive to  $\delta^{13}\text{C}_{\text{CH}_4}$  growing more negative. But that explanation required a very significant change in OH and the slow timing of OH's effect on  $\delta^{13}\text{C}_{\text{CH}_4}$  meant that the sink hypothesis could not explain the large and rapidly occurring inter-annual variations observed in  $\delta^{13}\text{C}_{\text{CH}_4}$ .

The analysis in Nisbet et al. (2016) suggested that for variation in the OH sink alone to be large enough to match the atmospheric data, significant changes in atmospheric chemistry would be seen in CO and non-methane hydrocarbons. However, that analysis did not consider trends and interannual variations in soil methanotrophy or in Cl removal, each of which have large isotopic fractionation effects: the extent that methane removal by Cl and soil methanotrophy can have impact on  $\delta^{13}\text{C}_{\text{CH}_4}$  fluctuations has not been considered hitherto. But, the recent increase in atmospheric methane would require total removal rates to have decreased by about 8%, which is close to the total removal by Cl and soil, thus the 'Cl and soil' explanation would still require some decrease in OH as well. That possibility has been considered here by keeping sources constant after 2002 and fitting the data with annual changes in each of OH, Cl and soil removal in each region. A tighter constraint is applied for OH than for the other removal processes.

In the analysis here, inter-annual changes in the methane budget start in 1998 when  $\delta^{13}\text{C}_{\text{CH}_4}$  data have global coverage. This comes after a 'spin-up' period of more than forty years to allow for the slow response of  $\delta^{13}\text{C}_{\text{CH}_4}$  to budget changes (Tans 1997), and initialisation of sources is made consistent with Lassey et al. (2007). Seasonal cycles for emissions and for their  $\delta^{13}\text{C}$  signatures are fitted to the data in each region but do not change over time. The dominant methane removal by OH is from Spivakovsky et al. (2000) and the highly fractionating removal by Cl is using the "FULL2" seasonal and zonal distribution from Hossaini et al. (2016), which is predominantly in the Northern Hemisphere and so quite different to that of Allan et al. (2007), used in Nisbet et al. (2016). The seasonal and zonal distribution for soil removal of methane and its fractionation effect are from Curry (2007) and Tyler et al. (2007); methane removal by exchange with the stratosphere is treated as removal in the low latitudes with an implicit source in the high latitudes consistent with Brewer Dobson circulation. Exchange between the four tropospheric regions is allowed to vary annually after 1998 but is kept consistent with an interhemispheric exchange rate of 1 – 1.2 years as used in other studies.

Total methane removal is consistent with the last IPCC assessment (Ciais et al., 2013), except that the new and more detailed treatment of Cl removal is half of what was given there. The methane budget derived for each region is summarized in Table 1 which shows the source analysis for the 2000 – 2005 period. The total sources estimated here and their attribution into four regions is sensitive to our use of the Spivakovsky et al. (2000) distribution for OH but, as that is well backed up by other studies, the budgets can be given at least a medium confidence level. Note that the relatively large amount of Cl removal in the 30-90°N given by Hossaini et al. (2016) makes the sources there significantly isotopically lighter than in the other regions.

Table 1. Average methane removal rates and source emissions derived in this analysis and changes. The upper two parts of the table summarise lifetimes and then sources with their  $\delta^{13}\text{C}_{\text{CH}_4}$  values, averaged over the 2000 – 2005 period, that match the observed mole fraction and  $\delta^{13}\text{C}_{\text{CH}_4}$  data. Mean lifetimes include cross-tropopause transport followed by removal in the stratosphere consistent with Ciais et al (2013). Larger emission uncertainties in most regions than in the global total reflect the uncertainties in exchange rates between the semi-hemispheres. The lower parts summarise changes in emissions and their  $\delta^{13}\text{C}_{\text{CH}_4}$ , or in the removal rate and its kinetic isotope effect, that match data over 2007-2014 and then 2015-2018.

	30-90°S	0-30°S	0-30°N	30-90°N	Global
OH lifetime (yr)	18.6	6.8	7.1	15.0	9.8
Cl lifetime (yr)	1445	763	281	182	361
Soil lifetime (yr)	655	132	165	125	173
Mean lifetime (yr)	17.9	5.7	5.9	12.5	8.3
Mean KIE	-5.4‰	-4.7‰	-5.3‰	-12.3‰	-6.2‰
Mean source over 2000 – 2005					
Tg/yr	62±2	173±4	198±4	161±4	594±2
annual cycle (Tg/yr)	47	31–	165	270	378
mean $\delta^{13}\text{C}$	-51.8±0.4‰	-52.0±0.2‰	-51.9±0.2‰	-56.3±0.2‰	-53.1±0.2‰
annual cycle $\delta^{13}\text{C}$	8.7‰	2.7‰	7.3‰	13.8‰	7.8‰
Changes in Sources					
2007 – 2014 relative to 2000 – 2005					
Tg/yr	+4.8 ± 1	+4.9 ± 2	11.7 ± 2	5.0 ± 2	26.4 ± 2
$\delta^{13}\text{C}$	-0.7 ± 0.2‰	-0.2 ± 0.1‰	+0.2 ± 0.1‰	-0.4 ± 0.1‰	-0.2 ± 0.1‰
2015 – 2018 relative to 2000 – 2005					
Tg/yr	+7.7 ± 1	+12.4 ± 3	+17.6 ± 3	+6.3 ± 2	+43.8 ± 2
$\delta^{13}\text{C}$	-0.5 ± 0.2‰	-0.2 ± 0.2‰	-0.1 ± 0.2‰	-0.7 ± 0.1‰	-0.3 ± 0.2‰
Changes in Removal					
2007 – 2014 relative to 2000-2005					
Tg/yr	-3.2 ± 0.6	-2.7 ± 1.1	-10.2 ± 1.2	-3.5 ± 1.2	-19.6 ± 2.1
Change in KIE	+0.3 ± 0.1‰	+0.2 ± 0.1‰	-0.1 ± 0.1‰	+0.40 ± 0.3‰	+0.2 ± 0.1‰
2015 – 2018 relative to 2000 – 2005					
Tg/yr	-5.7 ± 0.5	-12.9 ± 0.7	-15.8 ± 1.0	-7.5 ± 0.6	-41.7 ± 1.5
Change in KIE	+0.3 ± 0.1‰	+0.2 ± 0.1‰	0 ± 0.1‰	+1.4 ± 0.2‰	+0.5 ± 0.1‰

Fits to the mole fraction and  $\delta^{13}\text{C}_{\text{CH}_4}$  data are shown in Figure 4 and the corresponding source emissions and their  $\delta^{13}\text{C}_{\text{SRC}}$  in Figure 5. Supporting Information Figure SI 1 shows that the fitting process does follow seasonal cycles in the data, but running 12-month means are used here to make differences between the fits clearer.

Because of different seasonal cycles and fractionation effects for each removal process, the net fractionation also has a significant seasonal cycle and differs between the semi-hemispheres. The budget totals fall between the top down and bottom up budgets given in the Ciais et al. (2013) review for the last IPCC Working Group I report. The lower parts of Table 1 summarise the *changes in source emissions* or *changes in removal* rates required after 2006 to follow the more recent mole fraction and  $\delta^{13}\text{C}_{\text{CH}_4}$  data and these are shown in Figure 5.



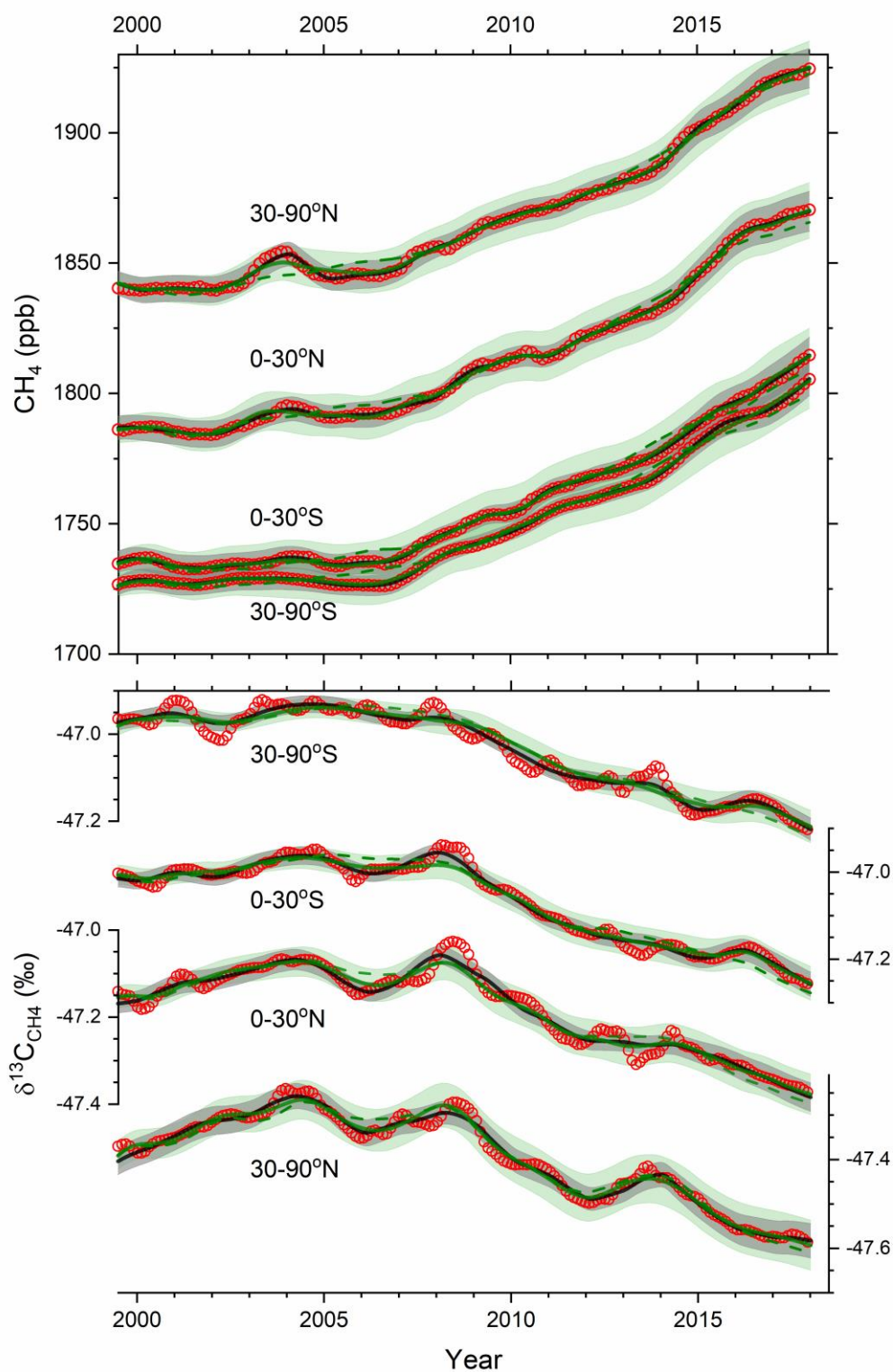


Figure 4. Running 12-month means for the  $\text{CH}_4$  mole fraction and  $\delta^{13}\text{C}_{\text{CH}_4}$  (red open circles) data are shown averaged over the four regions using data to mid-2018. Fits to the data are shown for “changes in source emissions” (black line and grey uncertainty range). Similar fits for “changes in removal” rate that have reduction in OH limited to 5% and in Cl and soil removal to 20%. are also shown (green lines and pale green uncertainty range). For comparison, dashed green lines show the results when OH variations are limited to 2%.

Fitting *changes in source emissions*, as shown in Figure 5, leads to global emissions increasing significantly in 2006 and again in 2013 and by about 20 Tg/yr in each case, but significant interannual variability is also seen in the semi-hemispheres. Sources have been getting isotopically lighter in 30-90°N and 0-30°S but not in 0-30°N. Note that relatively small emission in the 30-90°S zone makes it much more sensitive to inter-annual variations in transport, so the negative trend there is not robust. Anti-correlations between source emissions in 0-30°N and in 30-90°N such as seen over 2012-2018 also suggest that this simple budget analysis may not cover variations in the transport between regions.

Fits to the data with *changes in removal rate*, and sources constant after 2002, are also shown in Figure 5 (with more details in Figures SI 1 and SI 2 in Supporting Information). These have constraints on OH of  $\pm 5\%$  and on Cl and soil removal of  $\pm 20\%$ . As shown in Figure SI 1, these fits have removal rates starting near the high end of their allowed range over 2000 – 2005, dropping to an intermediate level in 2006, and then close to their lowest allowed level in 2013. This corresponds to transitions between three average removal rates corresponding to mean CH<sub>4</sub> lifetimes of 7.9, 8.3 and 8.6 years and, as shown in Figure 5, decreases in CH<sub>4</sub> removal are similar to optimised increases in emission when removal rates are fixed. Results from restricting OH changes to  $\pm 2\%$  are shown in Figures 4 and 5 and do not follow either mole fraction or  $\delta^{13}\text{C}_{\text{CH}_4}$  very well, while allowing OH changes of  $\pm 10\%$  made very little difference from  $\pm 5\%$ . When only OH is allowed to change, the fit to mole fraction is plausible, that for the  $\delta^{13}\text{C}_{\text{CH}_4}$  data is much worse.

Independent annual changes in removal for each of the four regions and for each of the OH, Cl and soil processes are highly variable as shown (Figure SI 2) implying that these are not well determined by the CH<sub>4</sub> data. Also, while large relative changes in Cl and soil removal together with smaller ones for OH can follow much of the trends and variability in mole fraction and  $\delta^{13}\text{C}_{\text{CH}_4}$ , large variations such as seen in the mole fraction in 2003 and in the  $\delta^{13}\text{C}_{\text{CH}_4}$  in 2008 are better reproduced by variations in sources.

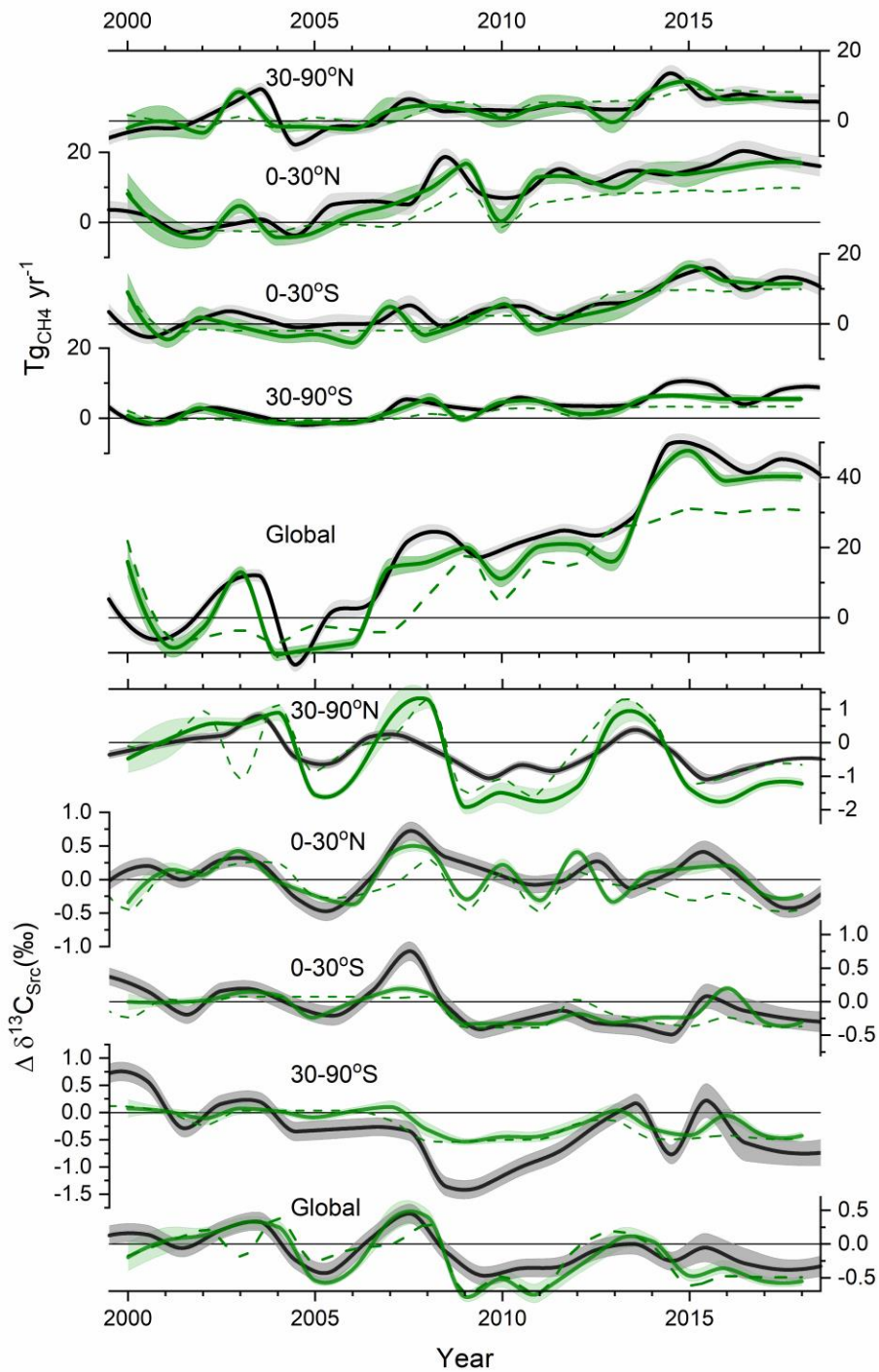


Figure 5. The upper panel shows interannual “changes in sources” fitted to the data for each region and the global average (black with grey uncertainty range) relative to 2000-2005. The alternative for “changes in removal” fitted to the data are also shown (green with pale green uncertainty range). Again, dashed green lines show results when OH variations are limited to 2%. The lower panel shows the changes in each region’s source  $\delta^{13}C$  value and the global average (black with grey uncertainty range) as well as changes in the net kinetic isotope effect (green with pale green uncertainty range) due to relative changes in the role of OH, Cl and soil.

Figure 4 shows that fits to the mole fraction and  $\delta^{13}\text{C}_{\text{CH}_4}$  data using *changes in removal rates* match the alternative explanation of *changes in source emissions* quite closely when large interannual changes in CI and soil removal are used with smaller ones for OH. However, large variations such as seen in the mole fraction in 2003 and in the  $\delta^{13}\text{C}_{\text{CH}_4}$  in 2008 are better reproduced by *changes in sources*. But this does not mean that removal rates have been constant. Rather, it shows that the problem is under-determined and that interannual fluctuations in soil methanotrophy and the CI sink could still be relevant, particularly in explaining some – but very unlikely, all – of the variations observed.

Answering the “sources vs sinks” debate clearly requires more isotopic data, especially in the tropics: to close the budget the enormous data gap in the tropical and sub-tropical regions needs to be filled. In particular,  $\delta^{13}\text{C}_{\text{CH}_4}$  is currently only measured at a very small number of stations. Ambiguities in the  $\delta^{13}\text{C}_{\text{CH}_4}$  budget analysis could be reduced by more extensive use of methane’s other isotopic parameters. In particular, long-term time series of  $\delta\text{D}_{\text{CH}_4}$  would provide powerful constraints on the problem, but currently data are minimal and there are very few active long-term measurement sites. Multiple substituted or ‘clumped’ isotopologues in methane also carry powerful information (Stolper et al., 2015) but this is frontier work only just beginning. Finding a good proxy for OH concentration in the atmosphere, and also determining the size of ongoing  $\text{CH}_3\text{CCl}_3$  emission is urgent.

A common factor linking many of the plausible explanations of the methane increase and isotopic shift since 2007 – whether growth in biogenic emissions, or sink declines – is that they may result from possible feedbacks to meteorological change that are not included in current climate models: the study by Naus et al. (2019) demonstrates the need for a proper modeling framework. The high growth in the Arctic in 2007, the global 2007-2009 growth, the impact of the 2010 La Niña in the tropics, and the four years of growth in 2014-2017 all suggest powerful feedbacks between the extreme meteorology of these years and methane growth.

## **8. Constraints on the Paris Agreement target, given methane’s growth.**

The methane rise has already had direct impact on the 2015 Paris Agreement, under which the UN Framework Convention on Climate Change (UNFCCC) set a goal of constraining future climate warming to a 2°C rise (UNFCCC 2015, 2017, 2018). Achievement of this ambitious warming outcome requires keeping the maximum anthropogenic radiative forcing to around  $3 \text{ W m}^{-2}$  (Meinshausen et al., 2011; Rogelj et al., 2012; see also IPCC 2018). This very challenging target demands rapid and severe cuts on anthropogenic greenhouse gas emissions, supported by an active program to monitor, quantify and control or remove them. The abatement potential of non- $\text{CO}_2$  reduction has a major impact on the allowable cumulative  $\text{CO}_2$  emission and cumulative abatement costs (Rogelj et al., 2015). In particular, to attain a sharp near-future reduction in radiative forcing by atmospheric greenhouse gases, the methane burden needs to be cut very rapidly indeed (Meinshausen et al., 2011; Rogelj et al., 2012).

The IPCC Summary for Policymakers on Global Warming of 1.5°C (IPCC 2018) updates the earlier Paris Agreement pathways, emphasising the importance of constraining global climate warming below a 1.5°C limit. This report highlights the major role that can be played by reductions in methane. Given sharp cuts in the atmospheric methane burden (as well as black carbon and  $\text{N}_2\text{O}$ ), coupled with strong reductions in  $\text{CO}_2$  emissions and carbon removal by forest regrowth, then it may yet be possible to attain the 1.5°C goal. However, if non- $\text{CO}_2$  emissions are not cut, then the only way of meeting the 1.5°C target is with heroic industrial-scale  $\text{CO}_2$  removal in addition to reforestation. The evidence for methane growth presented



here raises the possibility that should methane's current rate of increase continue for some decades, the Paris target may become intractable to attain.

Representative Concentration Pathway 2.6 (RCP2.6) (Moss et al., 2008; Meinshausen et al., 2011; van Vuuren et al., 2011; here defined as updated by Collins et al. 2013) is a standard emission scenario pathway that is capable of taking us to the Paris goal (Rogelj et al., 2012). RCP2.6 (Moss et al., 2008; Meinshausen et al., 2011; van Vuuren et al., 2011), defined as updated by Collins et al. (2013), can, if followed, achieve a Paris-compliant target, leading to likely peak warming between 1.2 – 2.0°C (Millar et al., 2017). But the RCP2.6 pathway is extremely demanding; moreover there are uncertainties over the extent of current warming (Cowtan et al., 2015; Schurer et al., 2018) and about the definition of the target (Hawkins et al., 2017).

A key contribution built into the RCP2.6 pathway is a 500 ppb reduction in methane this century, from 1754 ppb (taken as the baseline value in 2005) dropping after 2010 by about 6 ppb per year to 1452 ppb in 2050, then dropping by about 4 ppb a year to 1254 ppb in 2100 (Meinshausen et al., 2011). This would cut the total atmospheric methane burden by about 1400 Tg compared to the assumed baseline value, or by about 1660 Tg compared to the methane burden in 2018. To achieve this reduction would demand a cut in annual emissions of roughly 150 Tg during this century (2.77 Tg of CH<sub>4</sub>, distributed evenly into the global atmosphere, will increase the global mole fraction by 1 ppb).

But the methane burden is not dropping. The opposite is happening. By mid-2018, we were already >100 ppb above the methane projection used in RCP2.6 (Figure 6). If present growth rates are sustained to 2100, this would add another 600 ppb, taking methane to around 2400 ppb or nearly double the Paris-compliant IPCC RCP2.6 scenario. Even if methane growth only continued to 2050, the impact on compliance with Paris Agreement expectations would be severe. Figure 6 shows the discrepancy in warming impact ( $\text{mW m}^{-2}$ ) that has already opened up between the “Paris-compliant” RCP2.6 and the present (end-2017) methane burden. The methane increase since 2007 has given a radiative forcing of  $0.05 \text{ Wm}^{-2}$  for methane compared to RCP2.6. If this continues another 10 years it would be a radiative forcing of about  $0.1 \text{ Wm}^{-2}$ . That would be of the same order as the current increase in current CO<sub>2</sub> radiative forcing over 3 years.

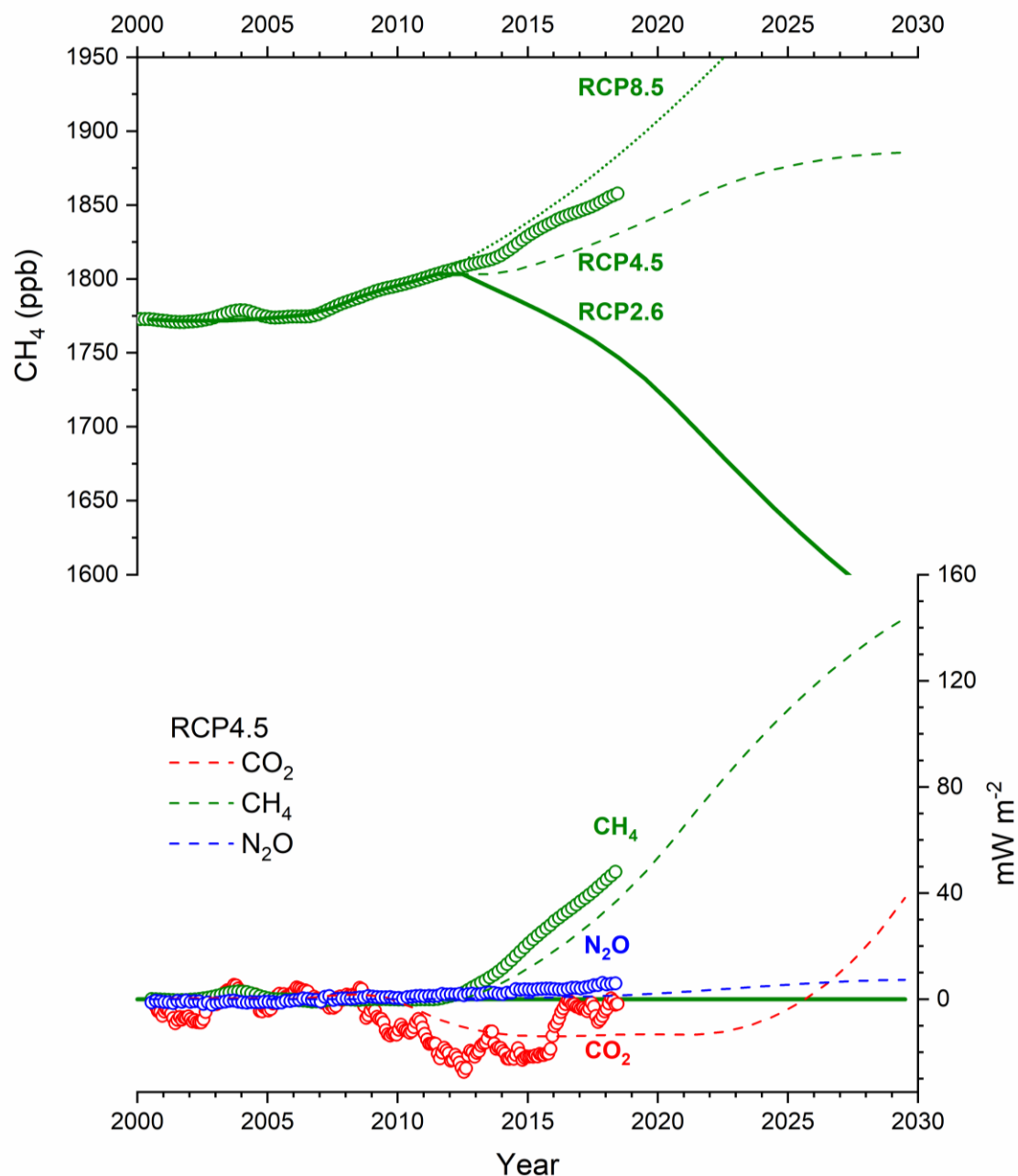


Figure 6. Top Panel: Observed global average methane (green open circles) at 3 monthly intervals (NOAA) are compared with the RCPs used in climate models (green solid, dashed and dotted lines).

Lower Panel shows differences in radiative forcing between actual evolution of methane (green open circles),  $\text{CO}_2$  (red circles) and nitrous oxide (blue circles) compared to the RCP2.6 pathway (0 solid line) and corresponding differences for the RCP4.5 pathway (dashed lines). Actual radiative forcing, 2000-present (shown as solid lines), from key gases in  $\text{mW m}^{-2}$  expressed as deviation from the RCP2.6 IPCC AR5 pathway (0 line on graph), which is compliant with the Paris target (Meinshausen et al., 2011; Rogelj et al., 2012). RCPs from IPCC AR5 Working Group 1 Appendix (Collins et al., 2013). For further context, the RCP4.5 scenario is also shown, to 2030 (dashed lines). For methane, current trends are higher even than the RCP4.5 scenario. Note that until the 2020s the RCP4.5 path does have lower  $\text{CO}_2$  concentrations than the RCP2.6 path.

There is a further troubling factor, not appreciated at the time the Paris Agreement was negotiated. Recent re-evaluation of the radiative forcing from methane (Etminan et al., 2016; Collins et al. 2018), including shortwave radiation effects, has found the forcing of methane since pre-industrial times to be about 25% stronger than the value used in the IPCC (Stocker et al., 2013) assessment. This re-evaluates methane's 100-year global warming potential (GWP<sub>100</sub>) factor as 32, or 14% higher than previously thought, though it should be noted that this value is provisional. GWP<sub>100</sub> is a measure of the integrated radiative forcing from equal-mass pulsed emission of a gas relative to CO<sub>2</sub>, and is used to determine trade-offs between different gases, evaluated over a 100-yr timescale. It is commonly used to determine how much CO<sub>2</sub> should be removed to compensate for methane emissions. But the evolving value for GWP<sub>100</sub> demonstrates the problems with using it to evaluate equivalences between different greenhouse gases. For the Kyoto Protocol the UNFCCC used a value of 21 from the IPCC's second assessment report, published in 1996, and a GWP<sub>100</sub> of 25 for the Doha Amendment covering 2013-2020 (<http://unfccc.int/resource/docs/2013/cop19/eng/10a03.pdf>)

The relevance of GWP<sub>100</sub> to the Paris Agreement has been discussed in recent papers (Fuglestad et al., 2018; Tanaka & O'Neill, 2018; Allen et al., 2018). CO<sub>2</sub>-equivalent emissions of methane calculated using GWP<sub>100</sub> do not predict the resultant warming, because methane does not remain in the atmosphere as long as CO<sub>2</sub>. Allen et al. (2018) show that in scenarios that achieve the Paris Agreement goals, GWP<sub>100</sub> misrepresents the impact of methane on temperature. They apply GWP<sub>100</sub> to a sustained change in the emission rate of methane (instead of a pulse), introducing a new parameter (GWP\*) to relate the impacts on radiative forcings of emissions of both short-lived and long-lived species. This method of relating CO<sub>2</sub>-equivalent emissions that are equivalent in their temperature response may allow relative evaluation of mitigation initiatives for different greenhouse gases. In contrast to GWP<sub>100</sub>, GWP\* correctly links increasing methane emissions with the resultant atmosphere warming, and decreasing emissions with atmospheric cooling.

In terms of radiative forcing, the methane growth in the past decade has added about 0.025 W m<sup>-2</sup> or 0.05 W m<sup>-2</sup> relative to the RCP2.6 expectation (Figure 6). At this growth rate, if sustained to 2100, methane would add more than 0.5 W m<sup>-2</sup> of unexpected forcing compared to the Paris target. Thus even if anthropogenic CO<sub>2</sub> emissions are successfully constrained to a RCP2.6-like pathway, the unexpected and sustained current rise in methane may so greatly overwhelm all progress from other reduction efforts that the Paris Agreement will fail.

In the context of the UN Framework Convention on Climate Change it should be noted the global methane emissions budget remains unconstrained (Kirschke et al., 2013; Saunio et al., 2016a,b) and there is concern about so-called “bottom up” inventory assessments, made by aggregating estimates of emissions from local, regional and national statistics and databases, total 600-900 Tg yr<sup>-1</sup>. In sharp contrast, if OH is buffered, “top-down” budgets calculated by modelling measurements of the ambient atmosphere, result in annual emissions between 560-580 Tg (Saunio et al., 2016a). One explanation of the discrepancy is that inventories may include double-counting of natural emissions (Thornton et al., 2016b). For the Paris Agreement to succeed and for reduction policies to be realistic, much better knowledge and much more accurate budgets are urgently required to bring “bottom-up” assessments into agreement with “top-down” measurements (Nisbet and Weiss, 2010).

## 9. Discussion

Speaking before the 2015 Paris Agreement, President Obama noted that in 2014 “the world economy grew while global carbon emissions from burning fossil fuels stayed flat. And what this means can't be overstated” (Obama, 2015). The Paris Agreement is ambitious, yet, as Obama pointed out, and despite the continuing very strong growth in the atmospheric burden

of CO<sub>2</sub>, there is optimism that reaching the target may just be feasible for CO<sub>2</sub> itself. But little attention was paid to the extraordinary rise in methane that year, despite methane's major contribution to the total anthropogenic warming to date. The Paris Agreement goals permit only a limited time to achieve net-zero CO<sub>2</sub> emissions before we overshoot the Agreement's warming limit. The more the atmospheric methane burden grows, the less time we have to reach net-zero CO<sub>2</sub> emissions.

The recent methane growth, and especially the acceleration in 2014, came as a surprise. If the processes driving growth are not directly anthropogenic, for example tropical rainfall or OH, Cl and soil sink changes, then the ~75 ppb scale of the recent rise so far is unprecedented in the Holocene record. For comparison, for fluctuations in the late Holocene Rhodes et al. (2017) find variabilities of 10-40 ppb. As noted above, large uncertainties remain in the global methane budget, and top-down and bottom-up estimates have not yet been reconciled. There remain major uncertainties about the emissions contributed by fossil fuel, agricultural and natural sources, and the proportions they constitute in the total budget. Unless the reasons for the current rise are understood, and it is determined whether or not growth is due to climate change feedbacks, it will not be possible to predict future trends in methane, nor to manage future feedbacks that may be driving the increase. Interestingly, although there may be no common causal link, the methane growth since 2007 and accelerated growth since 2014, both parallel increases in another major indicator of climate change, the rates of ice loss from both the Arctic (NSIDC 2018) and West Antarctica (IMBIE, 2018).

It is clear that if the Paris Agreement is to succeed, methane must be understood, but there are large unknowns. The key global deficiency, as noted by modelling studies (Turner et al., 2017; Rigby et al., 2017) is the lack of long-term measurement data from remote sites, especially in the tropics. There is urgent need for more *in situ* observations, from more locations, to constrain atmospheric inverse models of the methane budget. Isotopic measurement remains very sparse indeed. Satellite retrievals are unable to determine boundary layer methane abundance or to observe isotopologues accurately. To resolve the OH puzzle, more stations measuring long-term time series of methane mole fraction,  $\delta^{13}\text{C}_{\text{CH}_4}$  and  $\delta\text{D}_{\text{CH}_4}$  are needed, especially in the tropics, as well as a tighter understanding of ongoing methyl chloroform (CH<sub>3</sub>CCl<sub>3</sub>) emissions. But funding agencies prioritise hypothesis-testing and process-study campaigns over long-term monitoring.

There is a further ground for concern. There are very serious global implications for air quality if the rise in methane is due to a sharp drop in the oxidative capacity – the cleansing power – of the atmosphere, as implied by Turner et al. (2017) and to some extent by Rigby et al. (2017) and Poulter et al. (2017). If so, there are wide-ranging implications for many fields of atmospheric science that go far beyond greenhouse gas budget studies. It is very important to test the hypothesis that the oxidative capacity of the air has declined.

There is urgent need to quantify the necessary reduction of methane emissions to meet Paris Agreement goals, and to identify target sources for reduction. A key problem for the UNFCCC is the weakness of emissions inventories for methane sourced in so-called 'non-Annex 1' countries. These are emerging economies and low income "developing" nations, including China, India, Africa, and South America except French Guiana. These 'non-Annex 1' states include all tropical nations, including Singapore, Saudi Arabia and Qatar. The only tropical areas with more rigorous constraints on their inventory reporting are tropical territories of the US, Australia, France and the UK. Quantifying and regularly updating emissions from low-income tropical nations is urgent, both bottom-up and top-down. But setting up new monitoring sites is difficult as equipment import regulations are typically prohibitive, local interest and support may be low, and access difficult.

For methane, this is a grave weakness. Tropical and sub-tropical emissions are critical in the



global budget, both from anthropogenic sources including widely developing gasfields and some coalfields, the world's largest cattle populations and widespread deliberate biomass burning, and also from natural emissions in the very large wetlands of Amazonia and the Pantanal in South America, the Congo, Zambesi and other major river basins in Africa, and in S.E. Asia. Methane emissions have high societal cost (Shindell et al., 2017).

In contrast to CO<sub>2</sub>, methane provides many attractive, economically feasible and socially acceptable targets in reducing national and global emissions relatively inexpensively. A very attractive option is to be more vigorous in reducing fossil fuel emissions from gas leaks and coal processing. Emissions from fossil fuels and natural geological sources may be a higher fraction of the total global methane budget than previously thought, and may have been  $195 \pm 32 \text{ Tg yr}^{-1}$  in the 2003-2013 decade (Schwietzke et al., 2016). This result compares with a value of about 30.0 % of total emissions (CIRCA 160-170 TG) ESTIMATED FROM RADIOCARBON MEASUREMENTS (LASSEY ET AL, 2007). MOREOVER, ice-core studies suggest that natural emissions from geological sources during the Younger Dryas were no higher than 15.4 Tg annually (Petrenko et al. 2017), much less than the  $>50 \text{ Tg yr}^{-1}$  currently estimated. If geological emissions have not markedly changed since glacial times, modern annual anthropogenic fossil fuel emissions are correspondingly greater by  $\sim 35 \text{ Tg yr}^{-1}$ .

There are many good targets for reduction of methane emissions from fossil fuels. Leakage rates remain very significant (Peischl et al., 2016; Zavala-Araiza et al., 2015). Recent US evidence (Alvarez et al., 2018) suggests the gas leaks from the US supply chain are around 2.3% of gross gas production, or about  $13 \pm 2 \text{ Tg yr}^{-1}$ , which is  $\sim 60\%$  higher than US EPA inventory estimate. Recently, there has been rapid progress in technical methods to detect anthropogenic methane sources and in assessing their emission fluxes (Peischl et al., 2016; Zavala-Araiza et al., 2015; Schwietzke et al., 2017; Zazzeri et al., 2017). The disproportionate role of super-emitters has been widely observed (e.g. Zavala-Araiza et al., 2015). In the Barnett Shale gasfield in Texas, half of methane emissions came from 2% of the oil and gas facilities at any one time (Zavala-Araiza et al., 2015). Similar patterns of super-emission from abandoned wells and coal-mine vents were also found in Queensland and New South Wales, Australia (Zazzeri et al. 2016). Detection and monitoring of super-emitters is rapidly becoming simpler with the rapid advances in vehicle-based sensors (e.g. Zazzeri et al., 2015), aircraft measurement (Schwietzke et al., 2017) and satellite sensing, and continuous monitoring is now feasible around production sites, to watch super-emitters and potential super-emitters across their life cycle. Thus targeting super-emitters and working with operators to improve operational practices can enable rapid, low-cost cuts in emissions (Mayfield et al., 2017). Cutting gas leaks and landfill emissions, which is strongly supported by industry and the UN Climate and Clean Air Coalition (BP 2018), should be a global priority.

There are large potential benefits in dietary change (e.g. Ritchie et al. 2018), especially reduction in intense factory farming of cattle, but the net cut in methane emission from taking pastureland out of food production is not easily quantified. Replacing organic beef from rocky Scottish hills with soya grown intensively on former tropical rainforest or cerrado is not necessarily advantageous. Moreover forced reduction in cattle numbers is socially unacceptable in key regions: by far the largest population of ruminants is in India, while Ethiopia and South Sudan also have globally significant ruminant populations. Similarly, "mitigation" of wetland emissions has been suggested (e.g. Zhang et al. 2017). But such interventionist policies may not be feasible, could damage tropical wetlands with unpredictable consequences, and provoke strong local political antagonism to concerns about global warming.

There is more opportunity for cutting tropical biomass burning, which is easily observed and widely damaging agriculturally (exporting valuable soil nutrients to the wind). Reducing burning with targeted locally-based community-reward schemes is feasible, inexpensive, and should have wide beneficial impact. Removal of methane from ambient air has had little attention, but initial work suggests it may be simple and inexpensive in certain settings.

We may not be able to influence the factors driving the new rise in methane, especially if it is a climate change feedback, but by monitoring, quantifying and reducing the very large anthropogenic inputs, especially from the gas, coal and cattle industries, and perhaps by direct removal, we may be able to cut the total methane burden to be compliant with the Paris goals.

## **10. Conclusion**

The need to determine the factors behind the recent rise in methane is urgent: indeed, essential if global warming is to be limited within the Paris Agreement limits. If the main causes are increased anthropogenic emissions, they need to be reduced. If the increased methane burden is driven by increased emissions from natural sources, and if this is a climate feedback – the warming feeding the warming - then there is urgency to reduce anthropogenic emissions, which we can control. If, however, the increase in the methane burden is driven by a decline in the oxidative capacity of the atmosphere, and this is a climate feedback, then the implications are serious indeed.

Reducing methane emissions is feasible, especially from fossil fuel sources, and would have rapid impact on the global methane burden. This permits optimism but not complacency: the challenge is large. But there is no single silver bullet: there are many frontiers in methane research, and successfully meeting the Paris goals demands wide-ranging progress. Unless these questions are addressed, and much more attention paid to reducing methane emissions, especially from fossil fuels and biomass burning, the success of the Paris Agreement may be at risk.

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NOAA data are available from <https://www.esrl.noaa.gov/gmd/dv/data/>. RHUL data are being archived in the UK Centre for Environmental Data Analysis (CEDA) <http://catalogue.ceda.ac.uk/uuid/dd2b03d085c5494a8cbfc6b4b99ca702>.

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